

Appendix J, Spring Delta Outflow

Attachment J.3 Zooplankton-Delta Outflow Analysis

J.3.1 Model Overview

Zooplankton are an important food source for many larval, juvenile and small pelagic fishes in the Bay-Delta. Delta smelt and longfin smelt are two species that rely on zooplankton. This analysis followed the general framework of similar, prior analyses (Kimmerer 2002; Hennessy and Burris 2017; Greenwood 2018; California Department of Water Resources and U.S. Bureau of Reclamation 2023:2-10) to examine the relationship between Sacramento–San Joaquin Delta (Delta) outflow and Delta smelt and longfin smelt zooplankton prey density (catch per cubic meter) in the low salinity zone (i.e., 0.5–6 parts per thousand salinity). The analyses related prey density to Delta outflow during spring (March–May), summer (June–August), and fall (September–November) for the period from 2000 to 2021. This period generally represents the onset of the Pelagic Organism Decline ecological regime (Thomson et al. 2010). Zooplankton examined in the analyses were based on taxa (species or species groupings, split by life stage where appropriate) included in recent modeling and diet studies of both Delta smelt and longfin smelt (Slater and Baxter 2014; Smith 2021:45; Barros et al. 2022; Smith and Nobriga 2023). Results demonstrate the relationship between abundance of smelt prey in the Delta and outflow, controlled by CVP and SWP seasonal operations.

J.3.2 Model Development

J.3.2.1 Methods

Historical zooplankton data were synthesized using the R (R Core Team 2023) statistical software package *zooper* (Bashevkin et al. 2022; Bashevkin et al. 2023a, b). Data was subset as follows. For mysids, surveys included ‘EMP’ (Environmental Monitoring Program) data, whereas for other taxa surveys included ‘EMP’ as well as ‘20mm’ (20-mm Survey in spring), ‘STN’ (Summer Towntnet) and ‘FMWT’ (Fall Midwater Trawl). The data type chosen was ‘Community’, with size class of ‘Macro’ for mysids and ‘Micro’, ‘Meso’, and ‘Macro’ for other taxa. Only samples within the low salinity zone (salinity = 0.5–6 parts per thousand) were selected. The mean catch per unit effort (number per cubic meter) was calculated by year and for each season.

Historical Delta outflow data by year for each seasonal period were obtained from Dayflow via the Drought Data R package’s dataset `raw_hydro_1975_2022`.

For each taxon, mean annual loge-transformed catch per unit effort + 1 for each taxon was regressed against mean annual loge-transformed Delta outflow for each seasonal period. Statistically significant regressions (Table J.3-1 through Table J.3-3) were then applied to the 1922-2021 CalSim 3-modeled data for Baseline Conditions and Proposed Action scenarios, with predictions back-transformed to the original measurement scale (catch per unit effort, number per cubic meter) for summary of results.

Table J.3-1. Spring (March–May) zooplankton regression summary.

Taxon	Intercept	Slope	R ²	P
<i>Acartiella sinensis</i> (copepod) adults	8.134	-0.506	0.125	0.107
Cladocerans except <i>Daphnia</i>	-3.746	0.730	0.365	0.003
Copepod nauplii	10.145	0.140	0.079	0.205
Cyclopoid copepods except <i>Limnoithona</i> adults	7.441	0.042	0.005	0.764
<i>Daphnia</i> adults	-0.639	0.318	0.169	0.057
<i>Eurytemora affinis</i> (copepod) adults	0.234	0.528	0.255	0.016
Harpacticoid copepods	1.072	0.501	0.309	0.007
<i>Limnoithona</i> adults	7.973	0.135	0.030	0.439
Mysids	1.563	0.114	0.015	0.593
Other calanoid copepod adults	-1.593	0.669	0.210	0.032
Other calanoid copepod copepodites	2.296	0.469	0.357	0.003
<i>Pseudodiaptomus</i> (copepod) adults	4.496	0.053	0.001	0.874
<i>Pseudodiaptomus</i> (copepod) copepodites	0.882	0.476	0.149	0.076

Bolded text indicates statistically significant (P<0.05) regressions subsequently applied to CalSim 3-modeled data. Note: regressions were $\log_e(\text{mean annual catch per meter} + 1) = \log_e(\text{mean annual Delta outflow})$.

Table J.3-2. Summer (June–August) zooplankton regression summary.

Taxon	Intercept	Slope	R ²	P
<i>Acartiella sinensis</i> (copepod) adults	3.779	0.196	0.006	0.732
Cladocerans except <i>Daphnia</i>	11.625	-0.836	0.034	0.410
Copepod nauplii	10.883	0.163	0.038	0.385
Cyclopoid copepods except <i>Limnoithona</i> adults	9.204	-0.021	0.000	0.922
<i>Daphnia</i> adults	13.713	-1.316	0.104	0.143
<i>Eurytemora affinis</i> (copepod) adults	-2.567	0.445	0.055	0.294
Harpacticoid copepods	-2.539	0.788	0.119	0.117
<i>Limnoithona</i> adults	9.811	0.077	0.012	0.621
Mysids	0.065	0.364	0.031	0.211

Taxon	Intercept	Slope	R ²	P
Other calanoid copepod adults	8.603	-0.444	0.076	0.215
Other calanoid copepod copepodites	4.051	0.188	0.024	0.487
<i>Pseudodiaptomus</i> (copepod) adults	4.822	0.211	0.058	0.282
<i>Pseudodiaptomus</i> (copepod) copepodites	2.674	0.435	0.072	0.228

Regressions were $\log_e(\text{mean annual catch per meter}+1) = \log_e(\text{mean annual Delta outflow})$. None of the regressions were statistically significant ($P < 0.05$).

Table J.3-3. Fall (September–November) zooplankton regression summary.

Taxon	Intercept	Slope	R ²	P
<i>Acartiella sinensis</i> (copepod) adults	7.658	-0.119	0.005	0.752
Cladocerans except <i>Daphnia</i>	14.953	-1.375	0.053	0.300
Copepod nauplii	8.321	0.427	0.095	0.164
Cyclopoid copepods except <i>Limnoithona</i> adults	9.852	-0.069	0.002	0.862
<i>Daphnia</i> adults	6.854	-0.729	0.038	0.382
<i>Eurytemora affinis</i> (copepod) adults	-6.972	0.908	0.234	0.023
Harpacticoid copepods	4.114	0.054	0.000	0.960
<i>Limnoithona</i> adults	5.613	0.542	0.173	0.054
Mysids	-7.945	1.153	0.213	0.018
Other calanoid copepod adults	6.321	-0.436	0.012	0.621
Other calanoid copepod copepodites	2.286	0.359	0.032	0.426
<i>Pseudodiaptomus</i> (copepod) adults	11.444	-0.581	0.146	0.080
<i>Pseudodiaptomus</i> (copepod) copepodites	10.184	-0.484	0.047	0.334

Bolded text indicates statistically significant ($P < 0.05$) regressions subsequently applied to CalSim 3-modeled data. Regressions were $\log_e(\text{mean annual catch per meter}+1) = \log_e(\text{mean annual Delta outflow})$.

J.3.2.2 Assumptions/Uncertainty

This analysis is meant as a tool to compare mean abundance of zooplankton prey across different operation scenarios and is not a predictive tool.

While Delta outflow explains some of the variance in zooplankton CPUE, the relatively low R² values suggest that other factors contribute as well. Both Delta outflow and inflow are highly correlated (Kimmerer 2004); outflow is used as the variable of interest because it is more readily available and more easily linked to the position of X2. There are various hypothesized mechanisms for how flow affects zooplankton abundance in the low salinity zone via subsidies of zooplankton from higher abundance regions into the low salinity zone to subsidies of nutrients or phytoplankton (Hassrick et al. 2023; Kimmerer et al. 2019; Kimmerer 2002). A historical

regression of zooplankton CPUE with flow may be too simple and including other factors such as salinity, temperature, chlorophyll-*a*, residence time, etc. may have more explanatory power.

Historically, relationships between outflow and zooplankton abundance have changed over time (e.g., mysids in the summer) or relationships have become significant with increased outflow (e.g., *E. affinis* in the spring) or there is no relationship with outflow (e.g., rotifers and *E. affinis* in the summer) (Kimmerer 2009).

Zooplankton CPUE exhibits high variability, across regions, seasons, and years (Winder and Jassby 2011; Bollens et al. 2014; Lee et al. 2023) which may limit the statistical power to detect effects of flow alterations (Brandon et al. 2022).

J.3.2.3 Code and Data Repository

Biological data can be found online at www.wildlife.ca.gov/Conservation/Delta/Zooplankton-Study and on the ICF SharePoint in the Data and Code folder (Data and Code).

Hydrologic data can be found online at <http://www.water.ca.gov/dayflow/output/>

R code and results can be found on the ICF SharePoint in the Data and Code folder (Data and Code).

J.3.3 Results

Results are presented by taxon for statistically significant zooplankton regressions by water year type for each alternative in Table J.3-4 through Table J.3-17.

Tables include results from Explanatory 1 (EXP1), Explanatory 3 (EXP3), No Action Alternative (NAA), Alternative 2 with TUCPs (Alt2wTUCPwoVA), Alternative 2 without TUCPs (Alt2woTUCPwoVA), Alternative 2 with Delta VA (Alt2woTUCPDeltaVA), and Alternative 2 with systemwide VAs (Alt2woTUCPAIIVA).

Another set of tables include results from No Action Alternative (NAA), Alternative 1 (Alt1), Alternative 2 with TUCPs (Alt2wTUCPwoVA), Alternative 2 without TUCPs (Alt2woTUCPwoVA), Alternative 2 with Delta VA (Alt2woTUCPDeltaVA), and Alternative 2 with systemwide VAs (Alt2woTUCPAIIVA), Alternative 3 (Alt3), Alternative 4 (Alt4).

During spring months (March to May), the following taxa had a significant relationship with Delta outflow: Cladocerans (except *Daphnia*), *Eurytemora affinis* (copepod) adults, Harpacticoid copepods, Other calanoid copepod adults, and Other calanoid copepod copepodites (Table J.3-1).

J.3.3.1 Taxons

J.3.3.1.1 Cladocerans (except *Daphnia*)

For Cladocerans (except *Daphnia*) in the spring, during the **wet year type**, Alt3 had the highest CPUE (76) which was a 10% increase compared to the NAA. Alt1, Alt2wTUCPwoVA, Alt2woTUCPwoVA, and Alt4 had the lowest CPUE (68) which was a 1% decrease compared to the NAA. Alt2woTUCPAIIVA was no different from the NAA.

For **above normal year type**, Alt3 had the highest CPUE (54) which was a 12% increase compared to the NAA. Alt1 had the lowest CPUE (44) which was an 8% decrease compared to the NAA. Alt2wTUCPwoVA, Alt2woTUCPwoVA, and Alt4 all showed a 4% decrease compared to the NAA. Alt2woTUCPDeltaVA only showed a 2% decrease and Alt2woTUCPAIIVA was no different from the NAA.

For the **below normal year type**, Alt3 had the highest CPUE (36) which was a 12% increase compared to the NAA. Alt1 had the lowest CPUE (30) which was a 6% decrease compared to the NAA. Alt2wTUCPwoVA, Alt2woTUCPwoVA, and Alt4 all showed a 3% decrease compared to the NAA. Alt2woTUCPDeltaVA was no different from the NAA and Alt2woTUCPAIIVA showed a 6% increase.

For the **dry year type**, Alt3 had the highest CPUE (29) which was a 16% increase compared to the NAA. Alt1 had the lowest CPUE (24) which was a 4% decrease compared to the NAA. Alt2wTUCPwoVA, Alt2woTUCPwoVA, and Alt4 were no different than the NAA. Alt2woTUCPDeltaVA showed a 4% increase and Alt2woTUCPAIIVA showed an 8% increase.

For the **critical year type**, Alt2woTUCPAIIVA and Alt3 had the highest CPUE (18), which was a 20% increase compared to the NAA. Alt2woTUCPwoVA and Alt4 had the lowest CPUE (15) which was the same as the NAA. Alt1, Alt2woTUCPwoVA, and Alt2woTUCPDeltaVA showed a 13% increase.

Historically, in the low salinity zone (LSZ), CPUE of Cladocerans are lower when compared to more freshwater regions (Winder and Jassby 2011 Fig. 5, 7). While some marine and brackish water Cladocerans species are present in the San Francisco Estuary, freshwater Cladocerans tend to be more abundant in the Bay-Delta system. This could explain the lower CPUE observed during lower outflow scenarios and water year types.

Yet the mechanism for why CPUE increases in the low salinity zone during higher outflow has not been clearly and definitively established. Kimmerer (2002) found lower trophic level taxa (zooplankton) responded inconsistently with flow across seasons and historical periods. Kimmerer also found that chlorophyll showed little response to flow, suggesting a bottom up, “agricultural model” explanation for increased CPUE with higher flows is unlikely. Another possible mechanism is that increased flows also increase subsidies of zooplankton from higher abundance freshwater regions into the LSZ (Hassrick et al. 2023; Kimmerer et al. 2019).

Differences in operations during wet years appear to have little influence on the CPUE. Across all scenarios the CPUE remained unchanged or only decreased slightly (-1%) when compared to the NAA except for Alt3 which had increased outflow compared to all other scenarios.

As water year types became drier, the effect of scenarios on CPUE generally increased compared to the NAA. Alt1 generally had lower flows and lower CPUE for all year types, except for the critical year type. Similarly, Alt2wTUCPwoVA outflow and CPUE only increased for the critical year type. Alt2woTUCPwoVA outflow was only slightly less than the NAA for Wet, Above Normal and Below Normal years and almost the same for the Dry and Critical years. Alt2woTUCPDeltaVA only outflow increased compared to the NAA for Below Normal, Dry and Critical years. Alt2woTUCPAIIVA, outflow increased for Above Normal, Dry and Critical years. For all year types, Alt3 generally had the highest outflow and the largest CPUE increase compared to the NAA. Alt4 generally was either slightly below or above the NAA in terms of both outflow and CPUE percent difference.

For the critical year type all scenarios showed an increase or no change in CPUE from the NAA.

J.3.3.1.2 Adult *Eurytemora affinis*

For adult *Eurytemora affinis* in the spring, during the **wet year type**, Alt3 had the highest CPUE (432) which was a 7% increase compared to the NAA. Alt1 had the lowest CPUE (396) which was a 2% decrease compared to the NAA. For **above normal year type**, Alt3 had the highest CPUE (340) which was a 9% increase compared to the NAA. Alt1 had the lowest CPUE (297) which was a 5% decrease compared to the NAA. For the **below normal year type**, Alt3 had the highest CPUE (258) which was a 10% increase compared to the NAA. Alt1 had the lowest CPUE (223) which was a 5% decrease compared to the NAA. For the **dry year type**, Alt3 had the highest CPUE (218) which was a 9% increase compared to the NAA. Alt1 had the lowest CPUE (192) which was a 4% decrease compared to the NAA. For the **critical year type**, Alt3 had the highest CPUE (158), which was a 9% increase compared to the NAA. Alt2woTUCPwoVA had the lowest CPUE (141) which was the same compared to the NAA.

Kimmerer (2002) found a significant relationship between adult *E. affinis* and outflow. However, this relationship was only present after 1987, the post *Potamocorbula amurensis* invasion period which also coincided with a seven-fold decline in *E. affinis*.

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J.3.3.1.3 Harpacticoids

For harpacticoids in the spring, during the **wet year type**, Alt3 had the highest CPUE (736) which was a 7% increase compared to the NAA. Alt1 had the lowest CPUE (678) which was a 2% decrease compared to the NAA. For the **above normal year type**, Alt3 had the highest CPUE (587) which was a 9% increase compared to the NAA. Alt1 had the lowest CPUE (517) which was a 4% decrease compared to the NAA. For the **below normal year type**, Alt3 had the highest CPUE (452) which was a 10% increase compared to the NAA. Alt1 had the lowest CPUE (393) which was a 5% decrease compared to the NAA. For the **dry year type**, Alt3 had

the highest CPUE (385) which was an 8% increase compared to the NAA. Alt1 had the lowest CPUE (342) which was a 4% decrease compared to the NAA. For the **critical year type**, Alt3 had the highest CPUE (284), which was an 11% increase compared to the NAA. Alt2woTUCPwoVA and Alt4 had the lowest CPUE (256) which was a less than 0.5% decrease compared to the NAA.

Harpacticoids are not well studied in the Bay-Delta system but do sporadically show up in Delta smelt and longfin smelt diets (Burris et al. 2022; Slater et al. 2019; Slater and Baxter 2014; Nobriga 2002) and are a primary prey for common carp, Sacramento sucker, splittail (Feyrer et al. 2003), and smaller fish species such as mosquitofish, inland silversides, gobies (Glibert et al. 2011). Harpacticoids are present throughout the Bay-Delta system year-round (Ambler et al. 1985). Harpacticoids tend to be associated with benthic environments and may not be as readily available for consumption by Delta smelt and longfin smelt. There has been no previous research that observed a relationship between harpacticoids and flow.

The mechanism for why CPUE increases in the low salinity zone during higher outflow has not been clearly and definitively established. Kimmerer (2002) found lower trophic level taxa (zooplankton) responded inconsistently with flow across seasons and historical periods. Kimmerer also found that chlorophyll showed little response to flow, suggesting a bottom up, “agricultural model” explanation for increased CPUE with higher flows is unlikely. Another possible mechanism is that increased flows also increase subsidies of zooplankton from higher abundance freshwater seasons into the LSZ (Hassrick et al. 2023; Kimmerer et al. 2019).

J.3.3.1.4 Other adult calanoid copepods

For other adult calanoid copepods in the spring, during the **wet year type**, Alt3 had the highest CPUE (333) which was a 9% increase compared to the NAA. Alt1 had the lowest CPUE (300) which was a 2% decrease compared to the NAA. For **above normal year type**, Alt3 had the highest CPUE (244) which was an 11% increase compared to the NAA. Alt1 had the lowest CPUE (206) which was a 6% decrease compared to the NAA. For the **below normal year type**, Alt3 had the highest CPUE (172) which was a 13% increase compared to the NAA. Alt1 had the lowest CPUE (143) which was a 6% decrease compared to the NAA. For the **dry year type**, Alt3 had the highest CPUE (138) which was an 11% increase compared to the NAA. Alt1 had the lowest CPUE (118) which was a 5% decrease compared to the NAA. For the **critical year type**, Alt3 had the highest CPUE (92), which was a 15% increase compared to the NAA. Alt2woTUCPwoVA and Alt4 had the lowest CPUE (80) which was the same as the NAA.

The other calanoid copepods species included as part of “other adult calanoid copepods” were: *Acartia* spp., unidentified calanoids, *Sinocalanus doerrii*, *Tortanus* spp., and Diaptomidae. These species have been found in larval Delta smelt and longfin smelt diets (Burris et al. 2022; Slater et al. 2019; Slater and Baxter 2014; Nobriga 2002). Historically, since its introduction in 1978, *S. doerrii* was most abundant in the Suisun Bay region, with peak abundance during the spring to summer season, though it was noted that flow during winter would advect individuals into San Pablo Bay (Ambler et al. 1985). Since then, the range of *S. doerrii* has shifted landwards, likely because of effects of *Potamcorbula amurensis* grazing (Kimmerer et al. 1998). Therefore, increased flow is possibly advecting more individuals from more freshwater regions into the LSZ.

The mechanism for why CPUE increases in the low salinity zone during higher outflow has not been clearly and definitively established. Kimmerer (2002) found lower trophic level taxa (zooplankton) responded inconsistently with flow across seasons and historical periods. Kimmerer also found that chlorophyll showed little response to flow, suggesting a bottom up, “agricultural model” explanation for increased CPUE with higher flows is unlikely. Another possible mechanism is that increased flows also increase subsidies of zooplankton from higher abundance freshwater seasons into the LSZ (Hassrick et al. 2023; Kimmerer et al. 2019).

J.3.3.1.5 Other copepodite calanoid copepods

The other calanoid copepods species included as part of “other copepodite calanoid copepods” were: *Acartia* spp., *Acartiella* spp., unidentified calanoids, *Eurytemora affinis*, *Sinocalanus doerrii*, *Tortanus* spp., and Diaptomidae. These species have been found in larval Delta smelt and longfin smelt diets (Burris et al. 2022; Slater et al. 2019; Slater and Baxter 2014; Nobriga 2002).

For other copepodite calanoid copepods in the spring, during the **wet year type**, scenario Alt3 had the highest CPUE (1757) which was a 6% increase compared to the NAA. Alt1 had the lowest CPUE (1626) which was a 2% decrease compared to the NAA. For **above normal year type**, scenario Alt3 had the highest CPUE (1423) which was an 8% increase compared to the NAA. Alt1 had the lowest CPUE (1264) which was a 4% decrease compared to the NAA. For the **below normal year type**, scenario Alt3 had the highest CPUE (1116) which was a 9% increase compared to the NAA. Alt1 had the lowest CPUE (978) which was a 4% decrease compared to the NAA. For the **dry year type**, scenario Alt3 had the highest CPUE (960) which was an 8% increase compared to the NAA. Alt1 had the lowest CPUE (860) which was a 3% decrease compared to the NAA. For the **critical year type**, scenario Alt3 had the highest CPUE (723), which was an 11% increase compared to the NAA. Alt2woTUCPwoVA had the lowest CPUE (656) which was a less than 0.5% increase compared to the NAA.

The mechanism for why CPUE increases in the low salinity zone during higher outflow has not been clearly and definitively established. Kimmerer (2002) found lower trophic level taxa (zooplankton) responded inconsistently with flow across seasons and historical periods. Kimmerer also found that chlorophyll showed little response to flow, suggesting a bottom up, “agricultural model” explanation for increased CPUE with higher flows is unlikely. Another possible mechanism is that increased flows also increase subsidies of zooplankton from higher abundance freshwater seasons into the LSZ (Hassrick et al. 2023; Kimmerer et al. 2019).

J.3.3.2 Tables

Table J.3-4. Mean catch per unit effort (CPUE) for **Cladocerans (except *Daphnia*)** in **spring** by modeled scenario and water year type.

Water Year Type	EXP1	EXP3	NAA	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA
Wet	88	78	69	68	68	68	69
Above Normal	65	54	48	46	46	47	48
Below Normal	47	38	32	31	31	32	34
Dry	38	30	25	25	25	26	27
Critical	22	19	15	17	15	17	18

Values are rounded to the nearest integer.

Table J.3-5. Mean catch per unit effort (CPUE) for **Cladocerans (except *Daphnia*)** in **spring** by modeled scenario and water year type.

Water Year Type	NAA	Alt1	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA	Alt3	Alt4
Wet	69	68 (-1%)	68 (-1%)	68 (-1%)	68 (-1%)	69 (0%)	76 (10%)	68 (-1%)
Above Normal	48	44 (-8%)	46 (-4%)	46 (-4%)	47 (-2%)	48 (0%)	54 (12%)	46 (-4%)
Below Normal	32	30 (-6%)	31 (-3%)	31 (-3%)	32 (0%)	34 (6%)	36 (12%)	31 (-3%)
Dry	25	24 (-4%)	25 (0%)	25 (0%)	26 (4%)	27 (8%)	29 (16%)	25 (0%)
Critical	15	17 (13%)	17 (13%)	15 (0%)	17 (13%)	18 (20%)	18 (20%)	15 (0%)

Percentage values in parentheses indicate the difference between NAA and each alternative. Values and percent difference are rounded to the nearest integer.

Table J.3-6. Mean catch per unit effort (CPUE) for *E. affinis* adults in **spring** by modeled scenario and water year type.

Water Year Type	EXP1	EXP3	NAA	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA
Wet	480	438	404	399	399	400	401
Above Normal	388	343	312	304	306	311	316
Below Normal	307	264	234	229	230	235	243
Dry	264	226	200	198	198	202	209
Critical	180	163	141	153	141	153	157

Values are rounded to the nearest integer.

Table J.3-7. Mean CPUE for *E. affinis* adults in **spring** by modeled scenario and water year type.

Water Year Type	NAA	Alt1	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA	Alt3	Alt4
Wet	404	396 (-2%)	399 (-1%)	399 (-1%)	400 (-1%)	401 (-1%)	432 (7%)	398 (-1%)
Above Normal	312	297 (-5%)	304 (-3%)	306 (-2%)	311 (0%)	316 (1%)	340 (9%)	305 (-2%)
Below Normal	234	223 (-5%)	229 (-2%)	230 (-2%)	235 (0%)	243 (4%)	258 (10%)	229 (-2%)
Dry	200	192 (-4%)	198 (-1%)	198 (-1%)	202 (1%)	209 (4%)	218 (9%)	198 (-1%)
Critical	141	150 (6%)	153 (9%)	141 (0%)	153 (9%)	157 (11%)	158 (12%)	142 (1%)

Percentage values in parentheses indicate the difference between NAA and each alternative. Values and percent difference are rounded to the nearest integer.

Table J.3-8. Mean CPUE for **Harpacticoids** in **spring** by modeled scenario and water year type.

Water Year Type	EXP1	EXP3	NAA	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA
Wet	813	746	690	682	682	684	685
Above Normal	666	592	541	528	531	539	548
Below Normal	533	462	412	404	405	414	428
Dry	463	399	355	352	352	359	370
Critical	321	292	255	276	256	276	282

Values are rounded to the nearest integer.

Table J.3-9. Mean CPUE for **Harpacticoids** in **spring** by modeled scenario and water year type.

Water Year Type	NAA	Alt1	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA	Alt3	Alt4
Wet	690	678 (-2%)	682 (-1%)	682 (-1%)	684 (-1%)	685 (-1%)	736 (7%)	681 (-1%)
Above Normal	541	517 (-4%)	528 (-2%)	531 (-2%)	539 (0%)	548 (1%)	587 (9%)	531 (-2%)
Below Normal	412	393 (-5%)	404 (-2%)	405 (-2%)	414 (0%)	428 (4%)	452 (10%)	404 (-2%)
Dry	355	342 (-4%)	352 (-1%)	352 (-1%)	359 (1%)	370 (4%)	385 (8%)	352 (-1%)
Critical	255	270 (6%)	276 (8%)	256 (0%)	276 (8%)	282 (11%)	284 (11%)	256 (0%)

Percentage values in parentheses indicate the difference between NAA and each alternative. Values and percent difference are rounded to the nearest integer.

Table J.3-10. Mean CPUE for **other calanoid copepod adults** in **spring** by modeled scenario and water year type.

Water Year Type	EXP1	EXP3	NAA	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA
Wet	380	340	306	302	302	303	304
Above Normal	289	247	219	212	214	218	223
Below Normal	216	178	152	148	149	153	160
Dry	178	146	124	123	123	126	131
Critical	109	96	80	89	80	89	91

Values are rounded to the nearest integer.

Table J.3-11. Mean CPUE for **other calanoid copepod adults** in **spring** by modeled scenario and water year type.

Water Year Type	NAA	Alt1	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA	Alt3	Alt4
Wet	306	300 (-2%)	302 (-1%)	302 (-1%)	303 (-1%)	304 (-1%)	333 (9%)	301 (-2%)
Above Normal	219	206 (-6%)	212 (-3%)	214 (-2%)	218 (0%)	223 (2%)	244 (11%)	213 (-3%)
Below Normal	152	143 (-6%)	148 (-3%)	149 (-2%)	153 (1%)	160 (5%)	172 (13%)	148 (-3%)
Dry	124	118 (-5%)	123 (-1%)	123 (-1%)	126 (2%)	131 (6%)	138 (11%)	123 (-1%)
Critical	80	86 (8%)	89 (11%)	80 (0%)	89 (11%)	91 (14%)	92 (15%)	80 (0%)

Percentage values in parentheses indicate the difference between NAA and each alternative. Values and percent difference are rounded to the nearest integer.

Table J.3-12. Mean CPUE for **other calanoid copepod copepodites** in **spring** by modeled scenario and water year type.

Water Year Type	EXP1	EXP3	NAA	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA
Wet	1930	1780	1653	1635	1635	1641	1643
Above Normal	1602	1434	1319	1291	1297	1315	1336
Below Normal	1301	1138	1023	1005	1005	1027	1059
Dry	1139	991	890	883	883	900	925
Critical	808	742	653	704	656	704	719

Values are rounded to the nearest integer.

Table J.3-13. Mean CPUE for **other calanoid copepod copepodites** in **spring** by modeled scenario and water year type.

Water Year Type	NAA	Alt1	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA	Alt3	Alt4
Wet	1653	1626 (-2%)	1635 (-1%)	1635 (-1%)	1641 (-1%)	1643 (-1%)	1757 (6%)	1633 (-1%)
Above Normal	1319	1264 (-4%)	1291 (-2%)	1297 (-2%)	1315 (0%)	1336 (1%)	1423 (8%)	1295 (-2%)
Below Normal	1023	978 (-4%)	1005 (-2%)	1005 (-2%)	1027 (0%)	1059 (4%)	1116 (9%)	1004 (-2%)
Dry	890	860 (-3%)	883 (-1%)	883 (-1%)	900 (1%)	925 (4%)	960 (8%)	882 (-1%)
Critical	653	690 (6%)	704 (8%)	656 (0%)	704 (8%)	719 (10%)	723 (11%)	657 (1%)

Percentage values in parentheses indicate the difference between NAA and each alternative. Values and percent difference are rounded to the nearest integer.

During fall months (September to November), the CPUE of following taxon was significantly related to Delta outflow: adult *Eurytemora affinis* and mysids (Table J.3-3).

For adult *Eurytemora affinis* in the fall, across all scenarios, CPUE was very low (≤ 5). When CPUE was rounded to the nearest integer there was often no change compared to the NAA. During the **wet year type**, scenario Alt3 had the highest CPUE (4) which was a 33% increase compared to the NAA. Alt1 had the lowest CPUE (2) which was a 33% decrease compared to the NAA. For the **above normal year type**, scenario Alt2wTUCPwoVA, Alt2woTUCPwoVA, Alt2woTUCPDeltaVA, Alt2woTUCPAllVA, Alt3 and Alt4 had the highest CPUE (3) which was no different compared to the NAA. Alt1 had the lowest CPUE (1) which was a 67% decrease compared to the NAA. For the **below normal year type**, all scenarios were no different from the NAA, the CPUE was 1. For the **dry year type**, scenarios Alt2wTUCPwoVA, Alt2woTUCPwoVA, Alt2woTUCPDeltaVA, Alt2woTUCPAllVA, Alt3, and Alt4 had the highest CPUE (2) which was no different from the NAA. Alt1 had the lowest CPUE (1) which was a 50% decrease compared to the NAA. For the **critical year type**, all scenarios were no different from the NAA, the CPUE was 1. (Provide biological, ecological, and operational explanation for the observations)

For mysids in the fall, during the **wet year type**, scenario Alt3 had the highest CPUE (18) which was a 20% increase compared to the NAA. Alt1 had the lowest CPUE (11) which was a 27% decrease compared to the NAA. For **above normal year type**, scenario Alt3 had the highest CPUE (15) which was a 25% increase compared to the NAA. Alt1 had the lowest CPUE (7) which was a 42% decrease compared to the NAA. For the **below normal year type**, scenario Alt3 had the highest CPUE (7) which was a 17% increase compared to the NAA. Alt1 had the lowest CPUE (5) which was a 17% decrease compared to the NAA. For the **dry year type**, scenarios Alt2woTUCPAllVA and Alt3 had the highest CPUE (8) which was a 14% increase compared to the NAA. All other scenarios showed no difference from the NAA, the CPUE was 7. For the **critical year type**, all scenarios were no different from the NAA, the CPUE was 4. (Provide biological, ecological, and operational explanation for the observations)

Table J.3-14. Mean CPUE for *E. affinis* adults in fall by modeled scenario and water year type.

Water Year Type	EXP1	EXP3	NAA	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA
Wet	4	5	3	3	3	3	3
Above Normal	3	4	3	3	3	3	3
Below Normal	2	3	1	1	1	1	1
Dry	3	3	2	2	2	2	2
Critical	0	2	1	1	1	1	1

Values are rounded to the nearest integer.

Table J.3-15. Mean CPUE for *E. affinis* adults in fall by modeled scenario and water year type.

Water Year Type	NAA	Alt1	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA	Alt3	Alt4
Wet	3	2 (-33%)	3 (0%)	3 (0%)	3 (0%)	3 (0%)	4 (33%)	3 (0%)
Above Normal	3	1 (-67%)	3 (0%)	3 (0%)	3 (0%)	3 (0%)	3 (0%)	3 (0%)
Below Normal	1	1 (0%)	1 (0%)	1 (0%)	1 (0%)	1 (0%)	1 (0%)	1 (0%)
Dry	2	1 (-50%)	2 (0%)	2 (0%)	2 (0%)	2 (0%)	2 (0%)	2 (0%)
Critical	1	1 (0%)	1 (0%)	1 (0%)	1 (0%)	1 (0%)	1 (0%)	1 (0%)

Percentage values in parentheses indicate the difference between NAA and each alternative. Values and percent difference are rounded to the nearest integer.

Table J.3-16. Mean CPUE for **mysids** in **fall** by modeled scenario and water year type.

Water Year Type	EXP1	EXP3	NAA	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA
Wet	19	22	15	15	15	15	15
Above Normal	15	16	12	13	13	13	13
Below Normal	11	14	6	6	6	6	6
Dry	13	15	7	7	7	7	8
Critical	3	8	4	4	4	4	4

Values are rounded to the nearest integer.

Table J.3-17. Mean CPUE for **mysids** in **fall** by modeled scenario and water year type.

Water Year Type	NAA	Alt1	Alt2wTUCPwoVA	Alt2woTUCPwoVA	Alt2woTUCPDeltaVA	Alt2woTUCPAIIVA	Alt3	Alt4
Wet	15	11 (-27%)	15 (0%)	15 (0%)	15 (0%)	15 (0%)	18 (20%)	15 (0%)
Above Normal	12	7 (-42%)	13 (8%)	13 (8%)	13 (8%)	13 (8%)	15 (25%)	13 (8%)
Below Normal	6	5 (-17%)	6 (0%)	6 (0%)	6 (0%)	6 (0%)	7 (17%)	6 (0%)
Dry	7	7 (0%)	7 (0%)	7 (0%)	7 (0%)	8 (14%)	8 (14%)	7 (0%)
Critical	4	4 (0%)	4 (0%)	4 (0%)	4 (0%)	4 (0%)	4 (0%)	4 (0%)

Percentage values in parentheses indicate the difference between NAA and each alternative. Values and percent difference are rounded to the nearest integer.

J.3.3.3 Figures

Spring Zooplankton

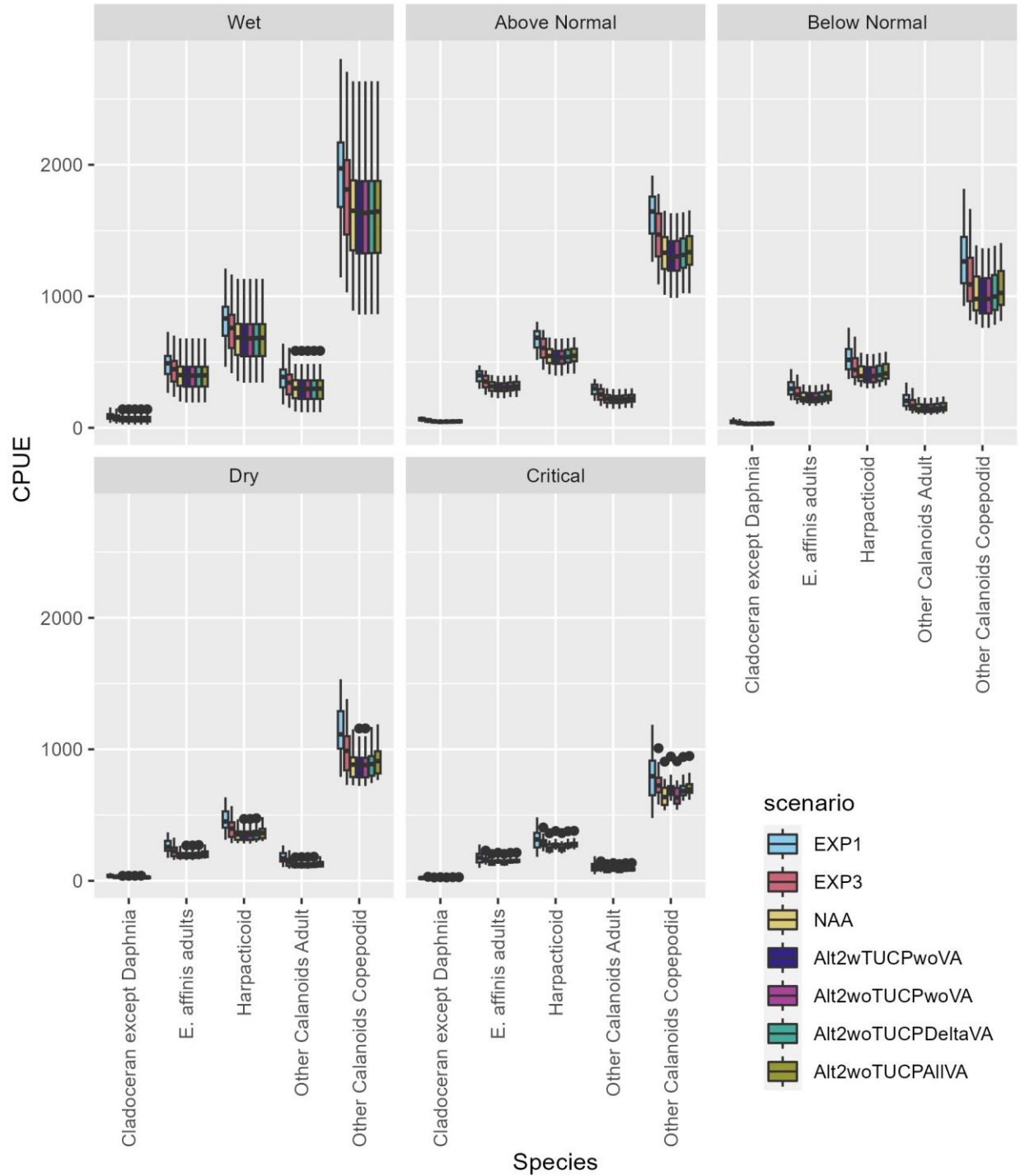


Figure J.3-1. Box Plots of CPUE of significant zooplankton species by scenario across different water year types for spring.

Spring Zooplankton

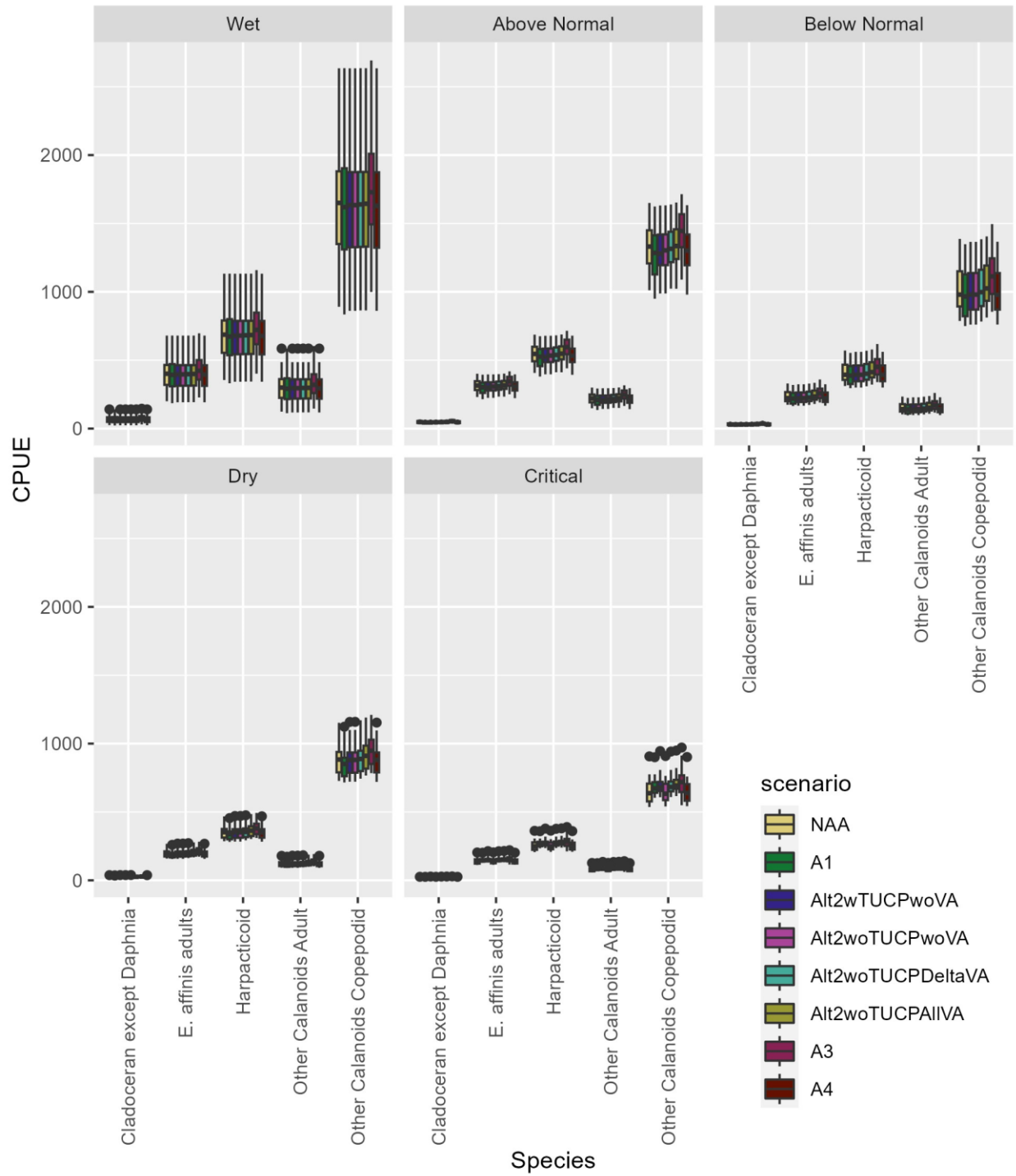


Figure J.3-2. Box Plots of CPUE of significant zooplankton species by scenario across different water year types for spring.

Cladoceran (except Daphnia) -Spring

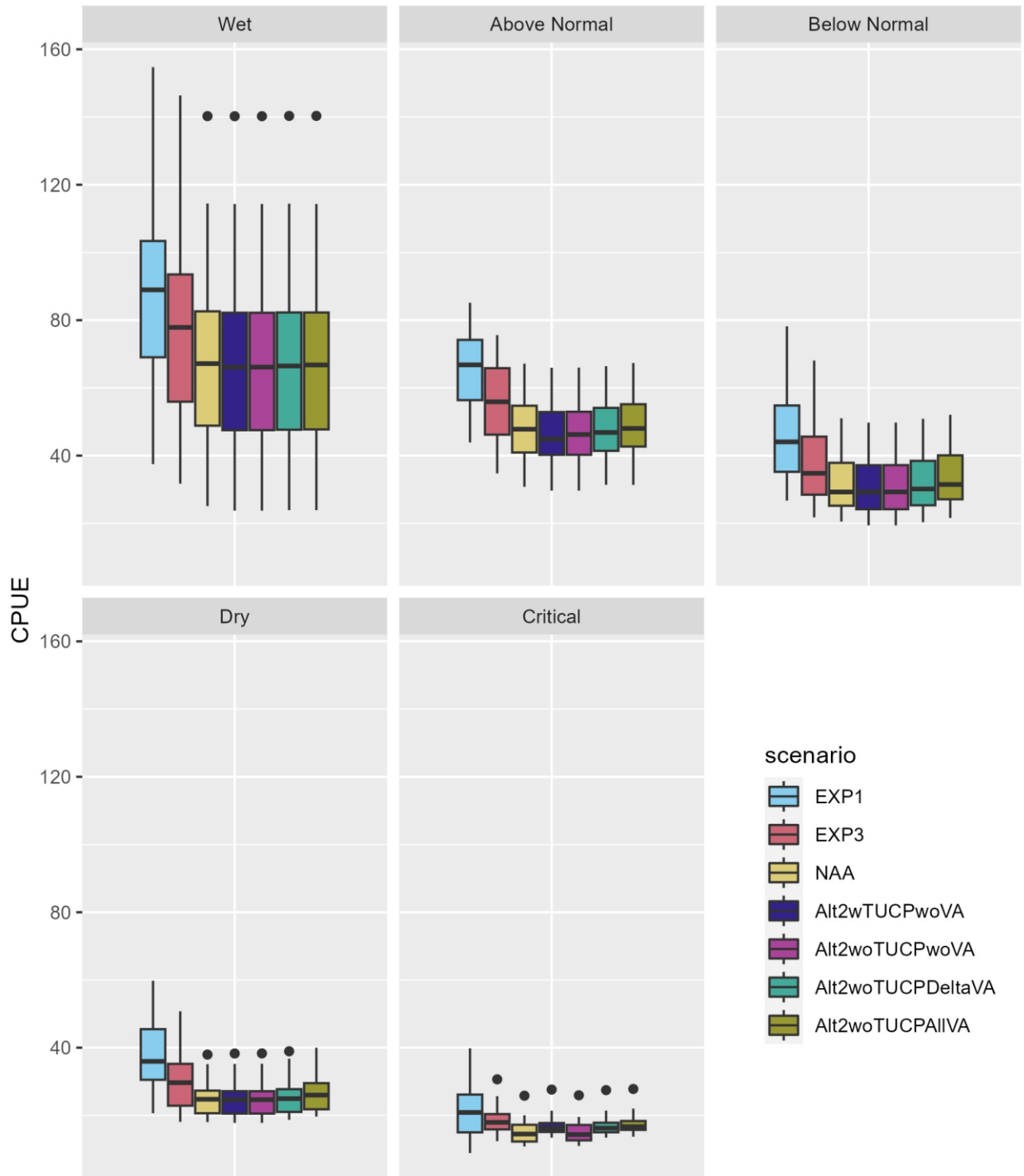


Figure J.3-3. Box Plots of CPUE of Cladocerans (except *Daphnia*) by scenario across different water year types for spring.

Cladoceran (except Daphnia) -Spring

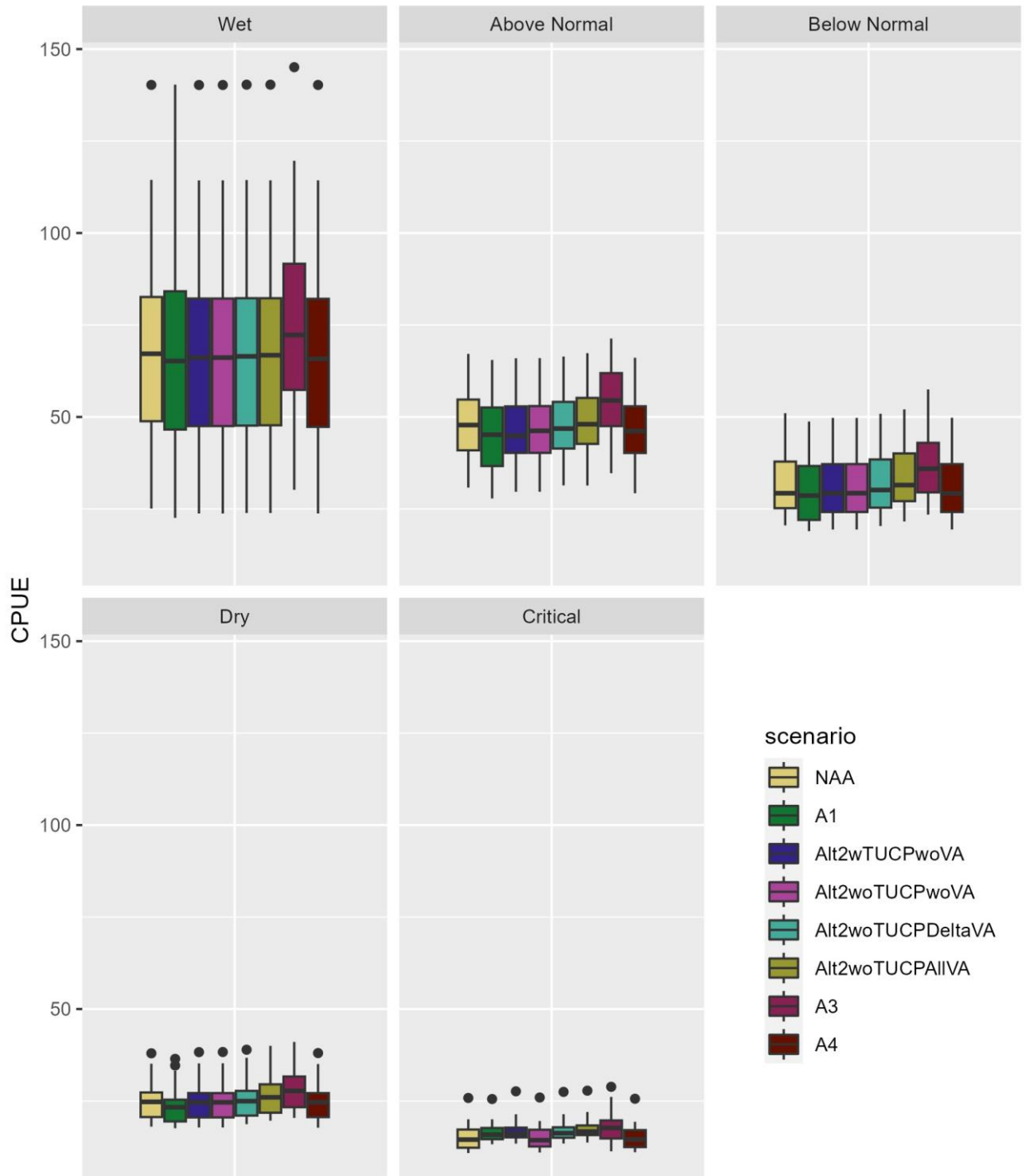


Figure J.3-4. Box Plots of CPUE of Cladocerans (except *Daphnia*) by scenario across different water year types for spring.

E. affinis adults -Spring

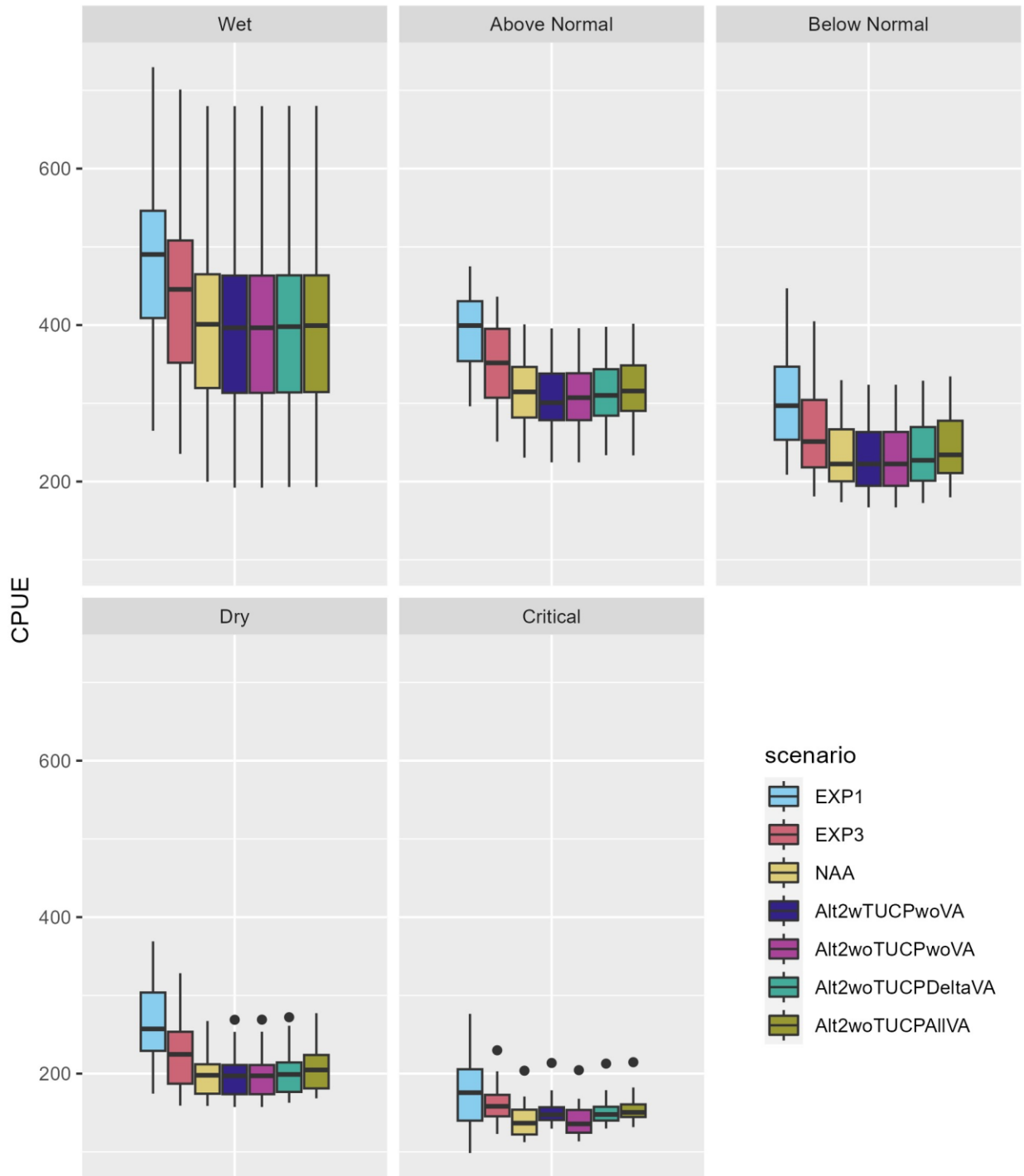


Figure J.3-5. Box Plots of CPUE of *E. affinis* adults by scenario across different water year types for spring.

E. affinis adults -Spring

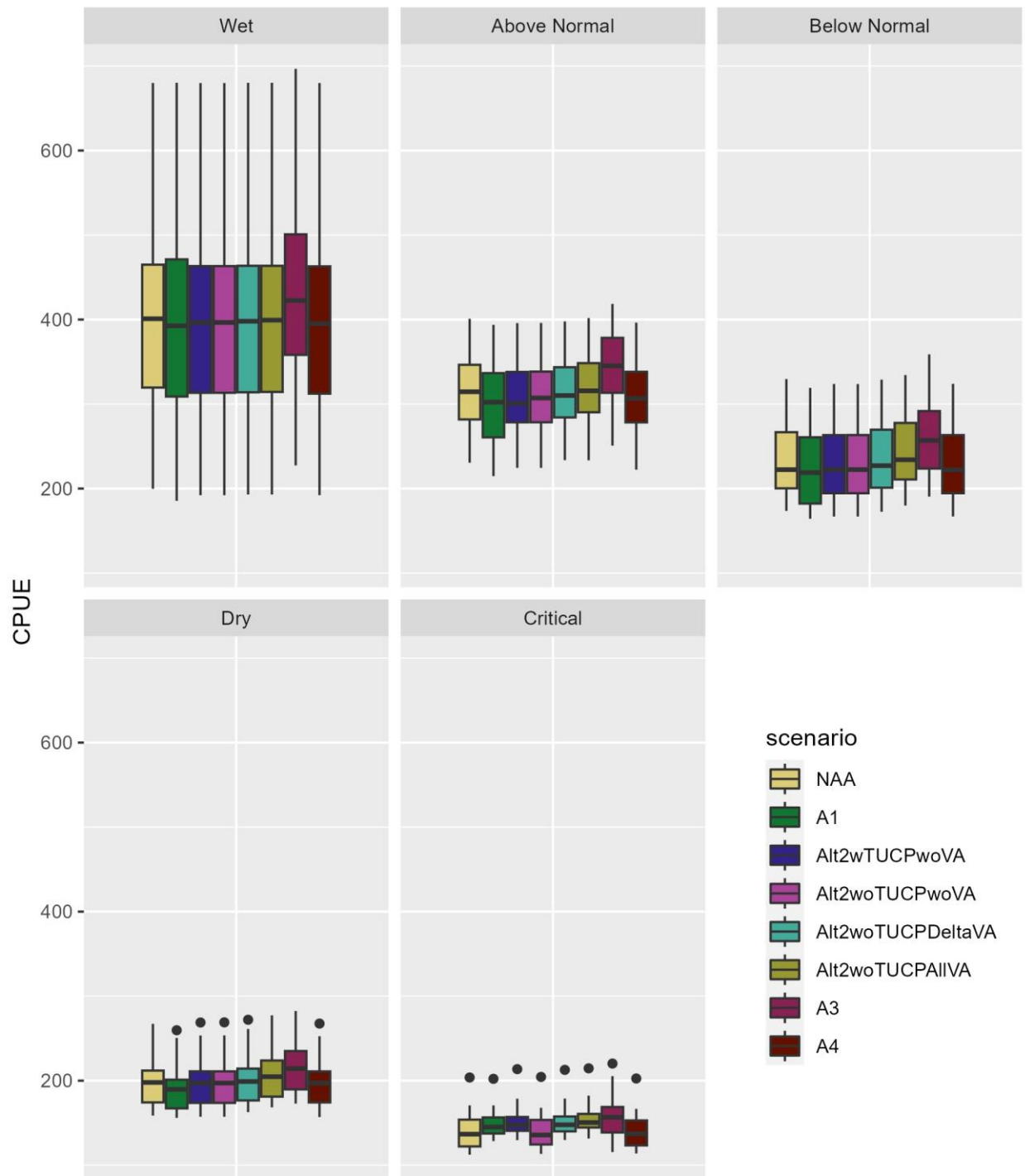


Figure J.3-6. Box Plots of CPUE of *E. affinis* adults by scenario across different water year types for spring.

Harpacticoid -Spring

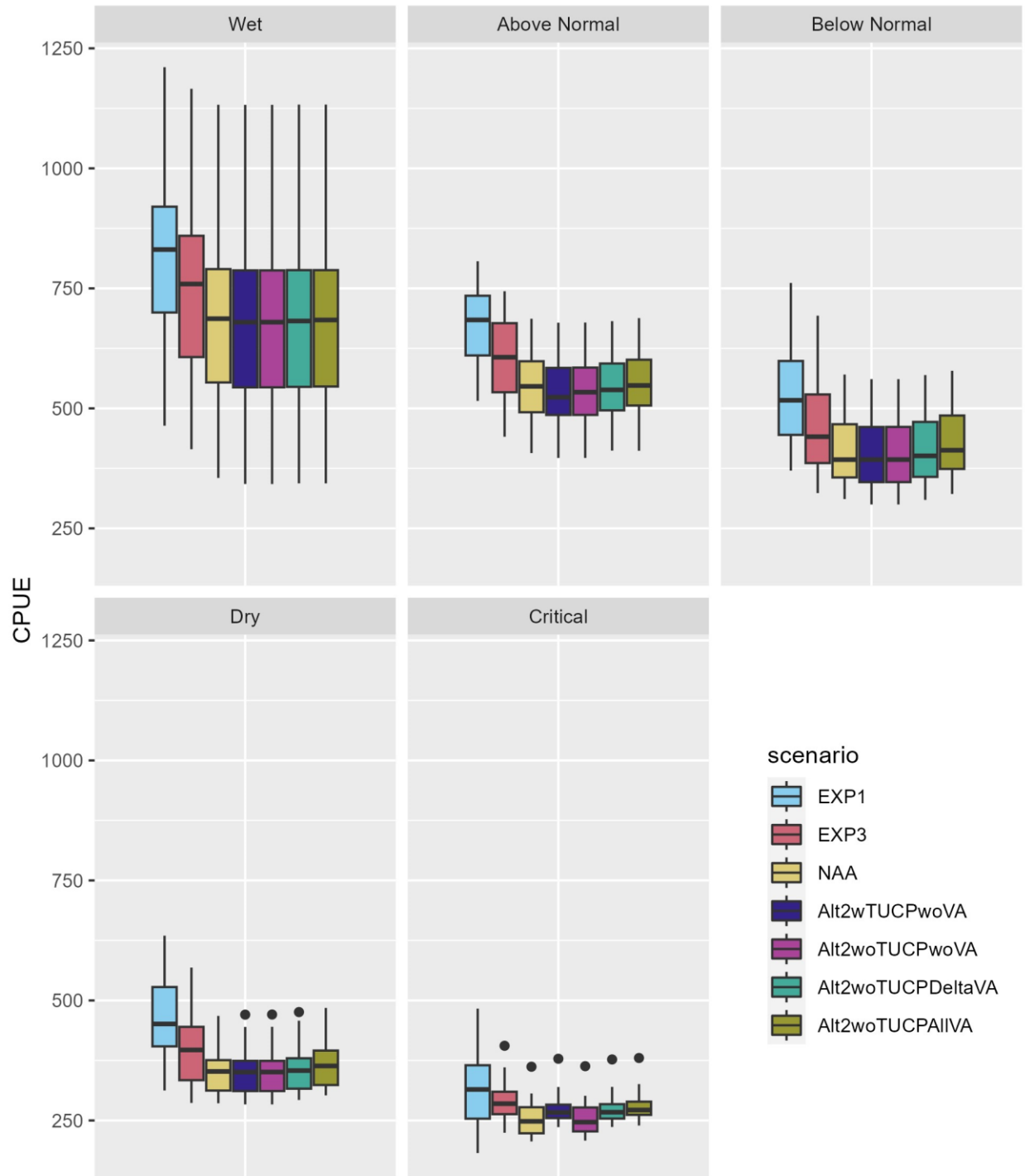


Figure J.3-7. Box Plots of CPUE of harpacticoids by scenario across different water year types for spring.

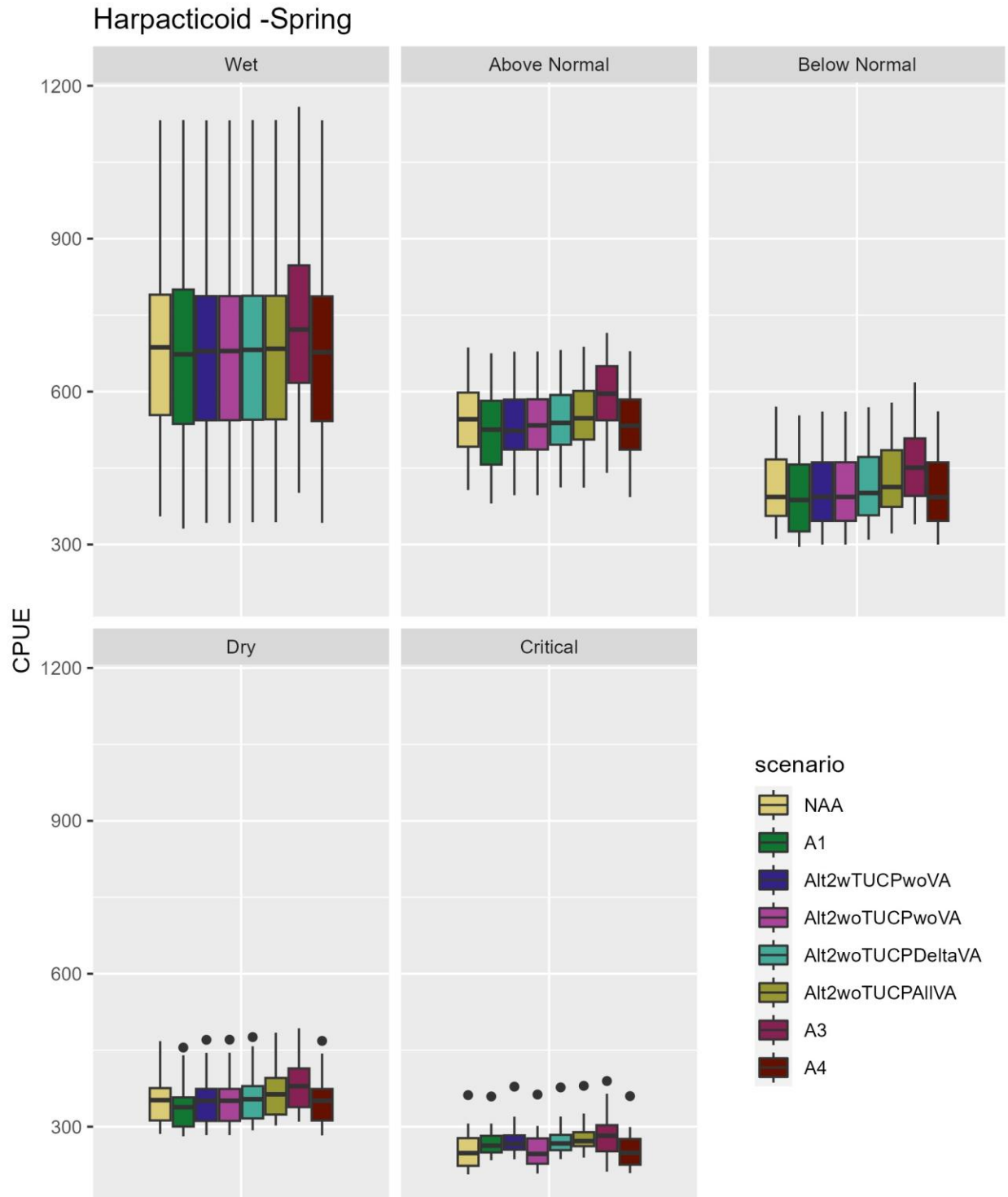


Figure J.3-8. Box Plots of CPUE of harpacticoids by scenario across different water year types for spring.

Other Calanoids Adult -Spring

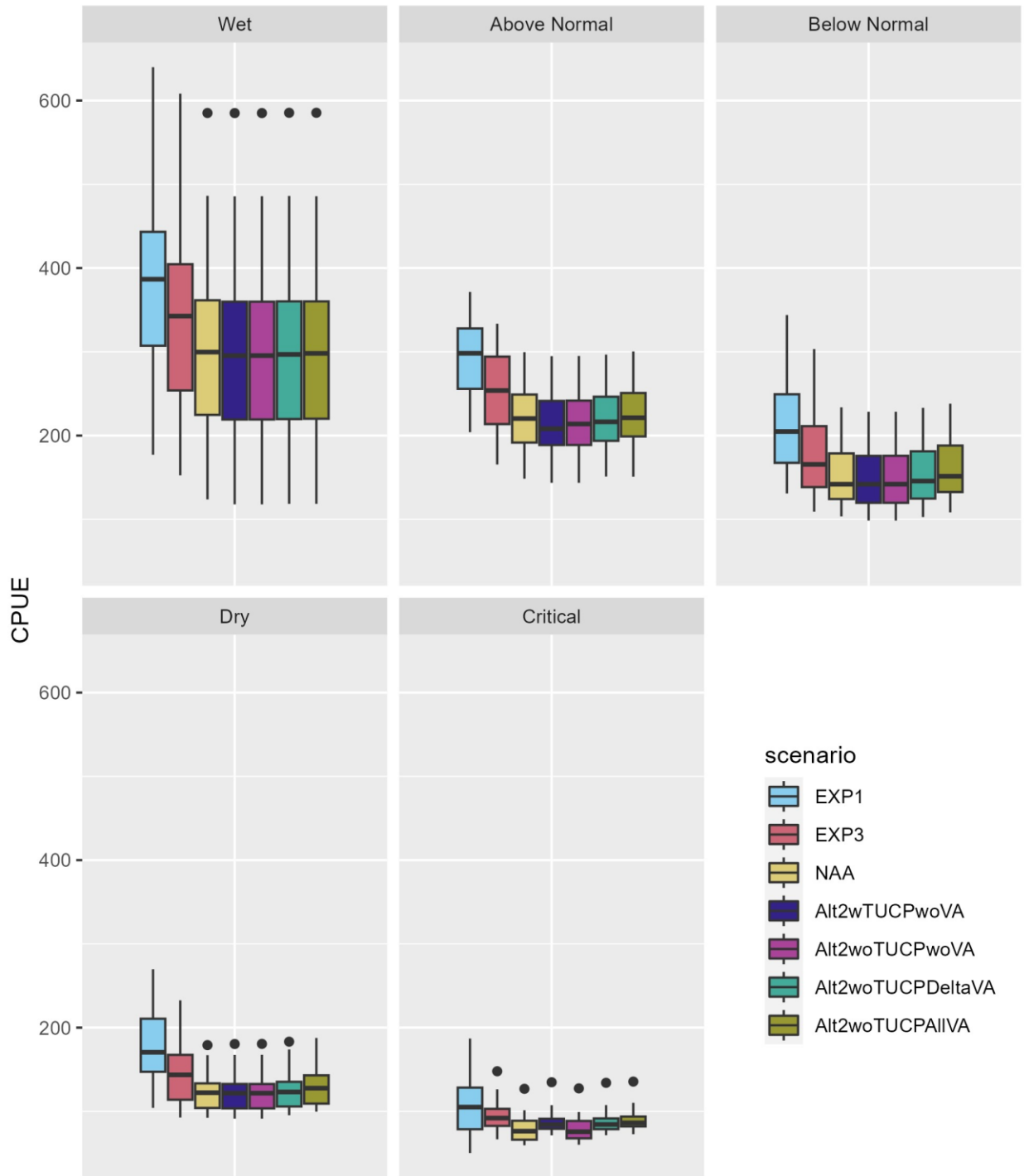


Figure J.3-9. Box Plots of CPUE of other calanoids adults by scenario across different water year types for spring.

Other Calanoids Adult -Spring

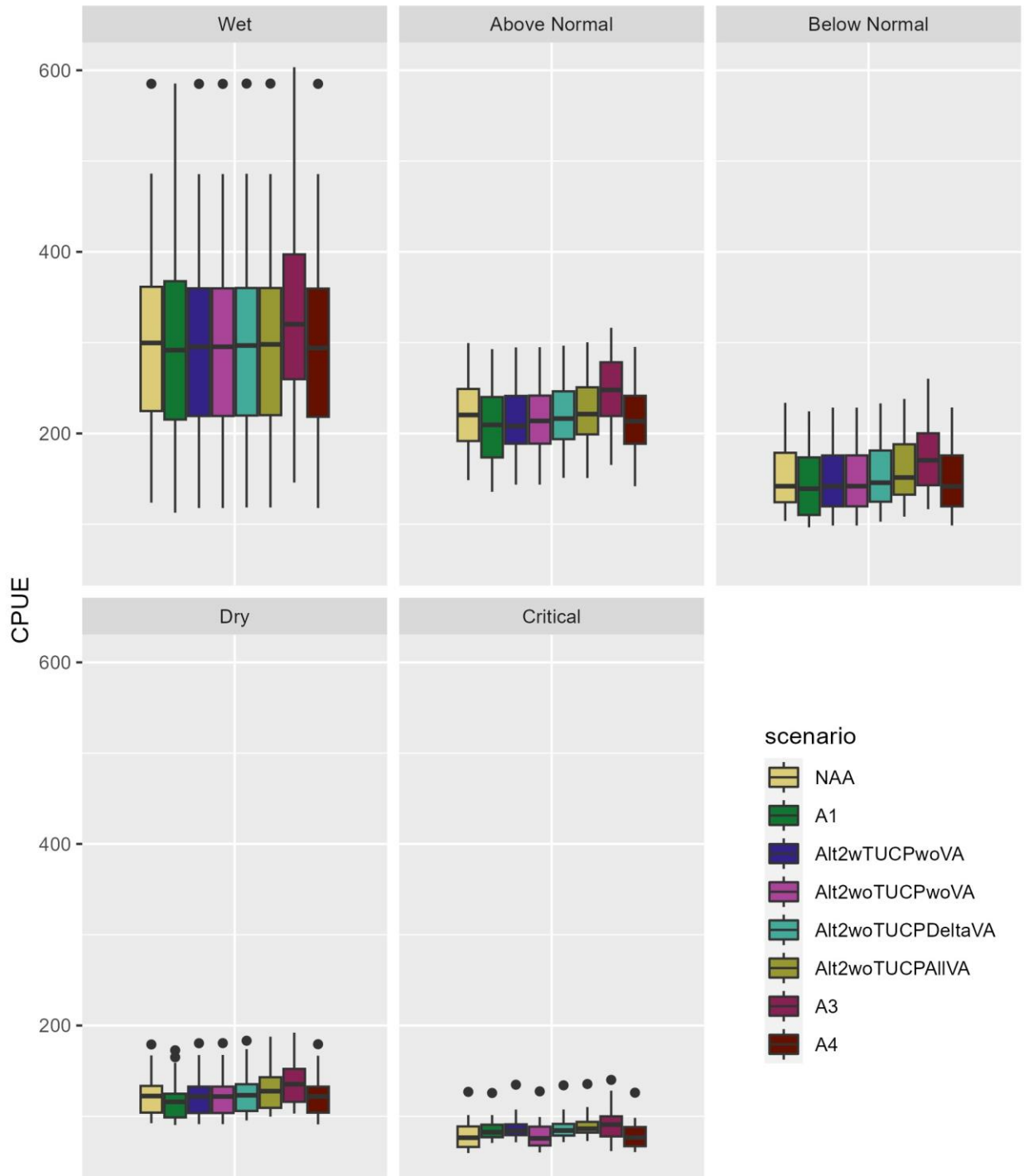


Figure J.3-10. Box Plots of CPUE of other calanoids adults by scenario across different water year types for spring.

Other Calanoids Copepodid -Spring

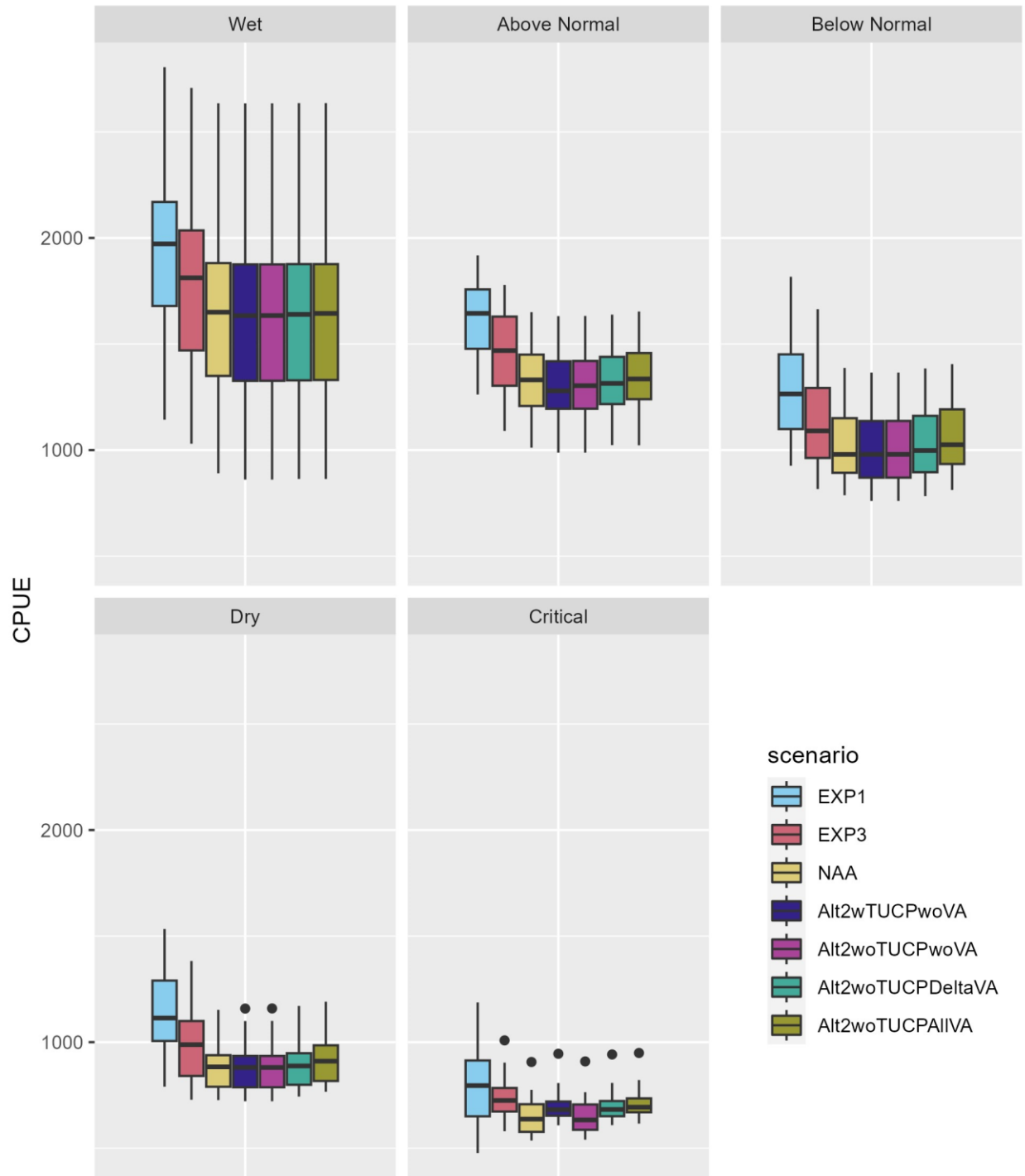


Figure J.3-11. Box Plots of CPUE of other calanoids copepodids by scenario across different water year types for spring.

Other Calanoids Copepodid -Spring

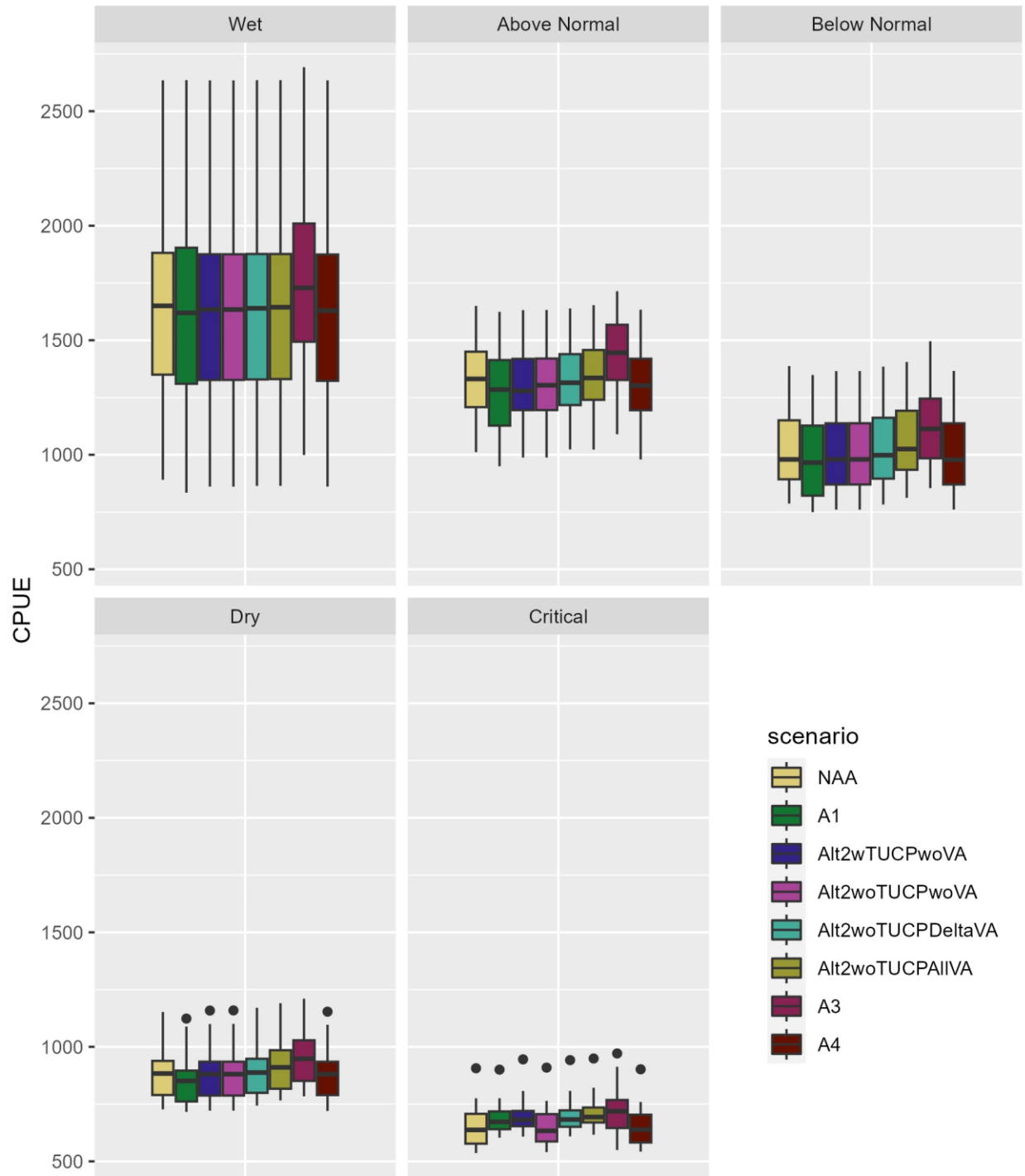


Figure J.3-12. Box Plots of CPUE of other calanoids copepodids by scenario across different water year types for spring.

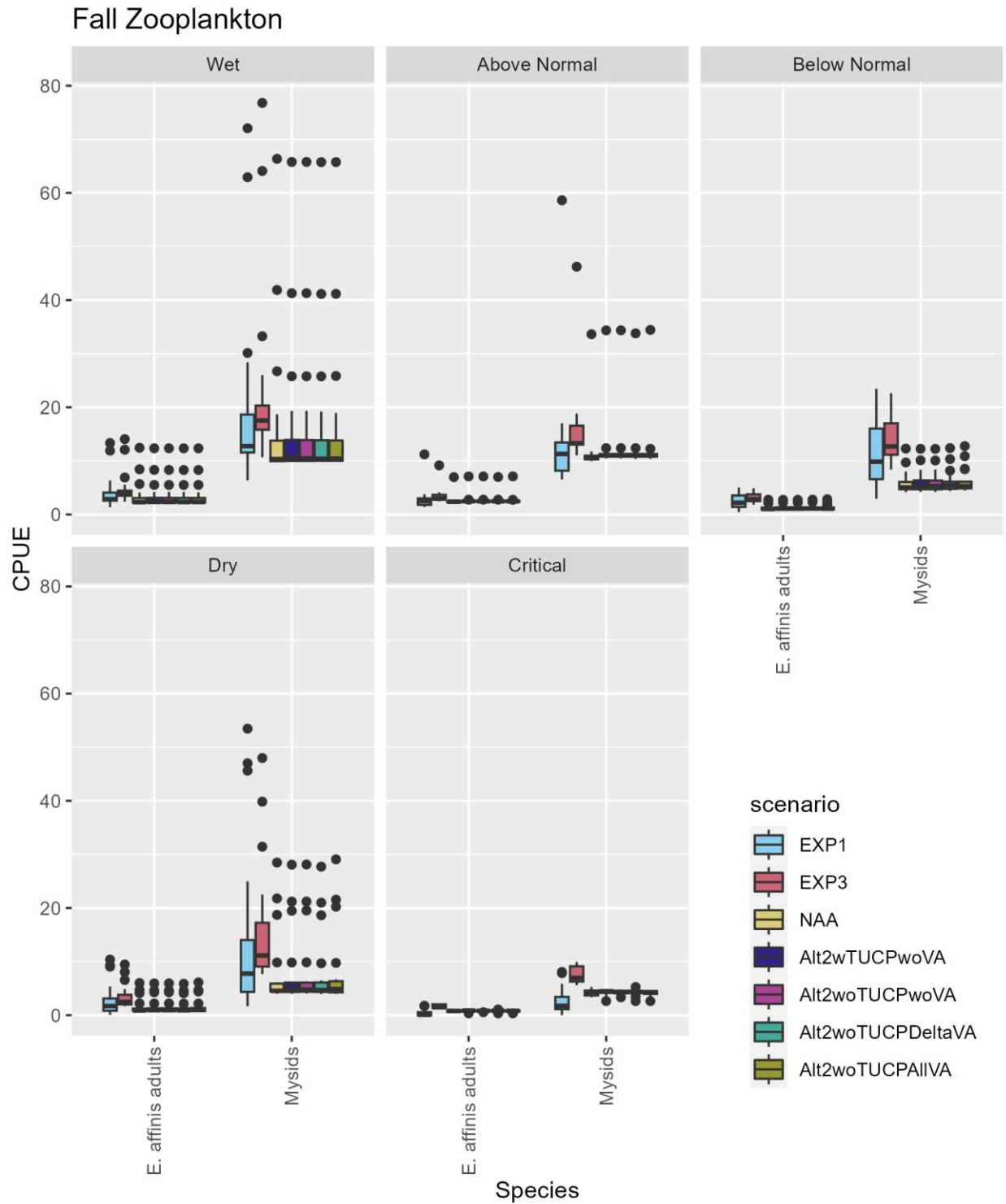


Figure J.3-13. Box Plots of CPUE of significant zooplankton species by scenario across different water year types for fall.

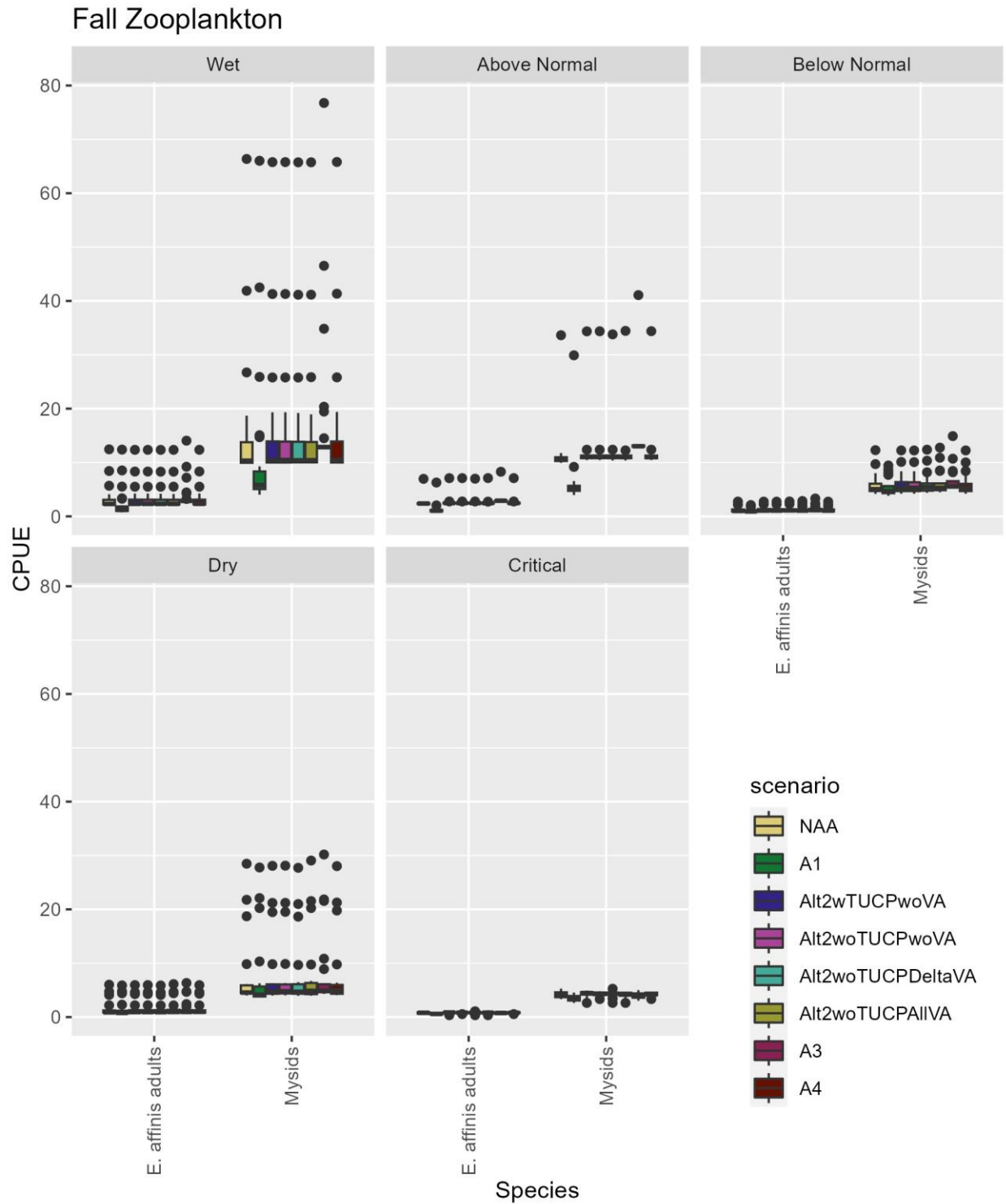


Figure J.3-14. Box Plots of CPUE of significant zooplankton species by scenario across different water year types for fall.

E. affinis adults -Fall

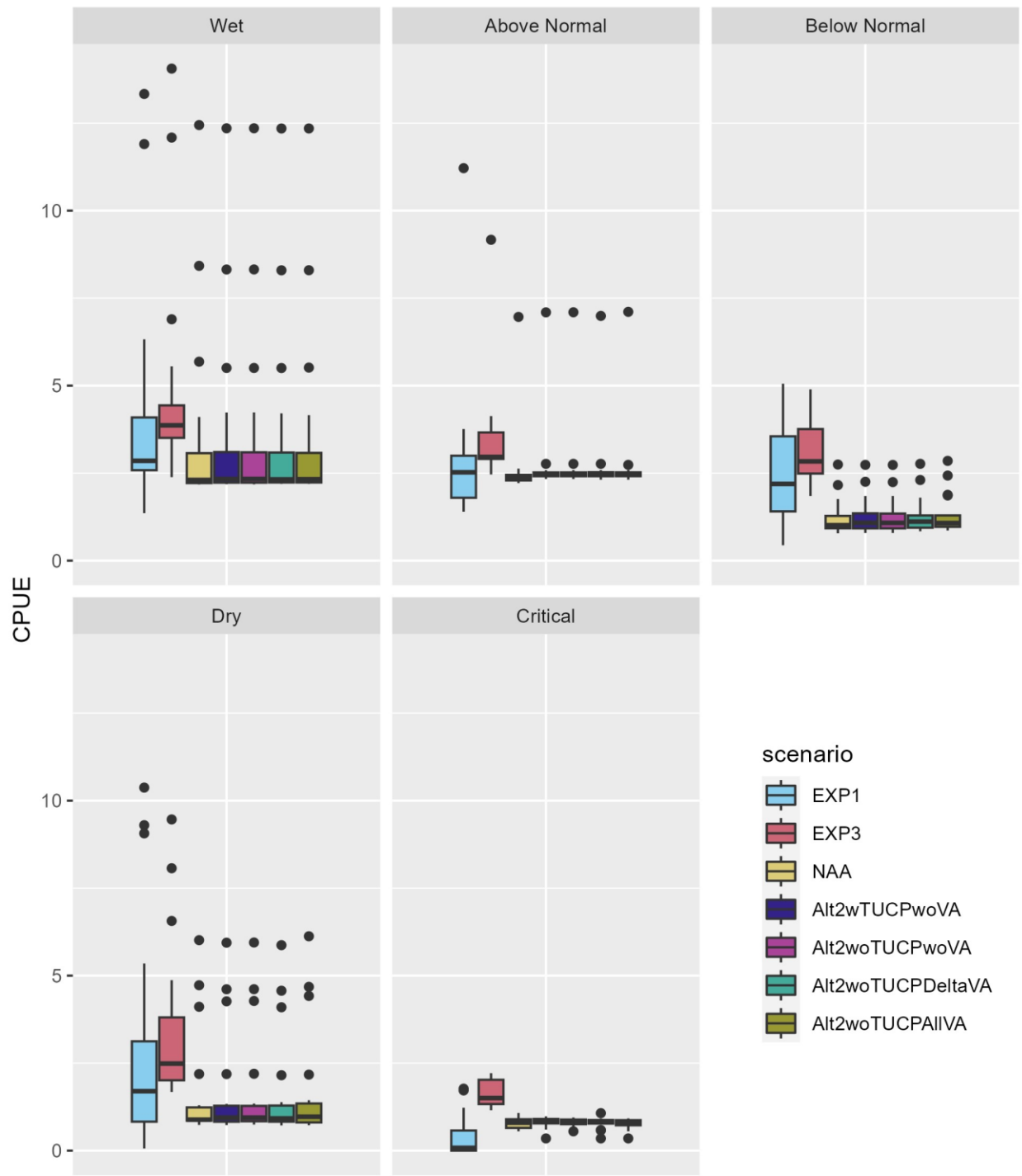


Figure J.3-15. Box Plots of CPUE of *E. affinis* adults by scenario across different water year types for fall.

E. affinis adults -Fall

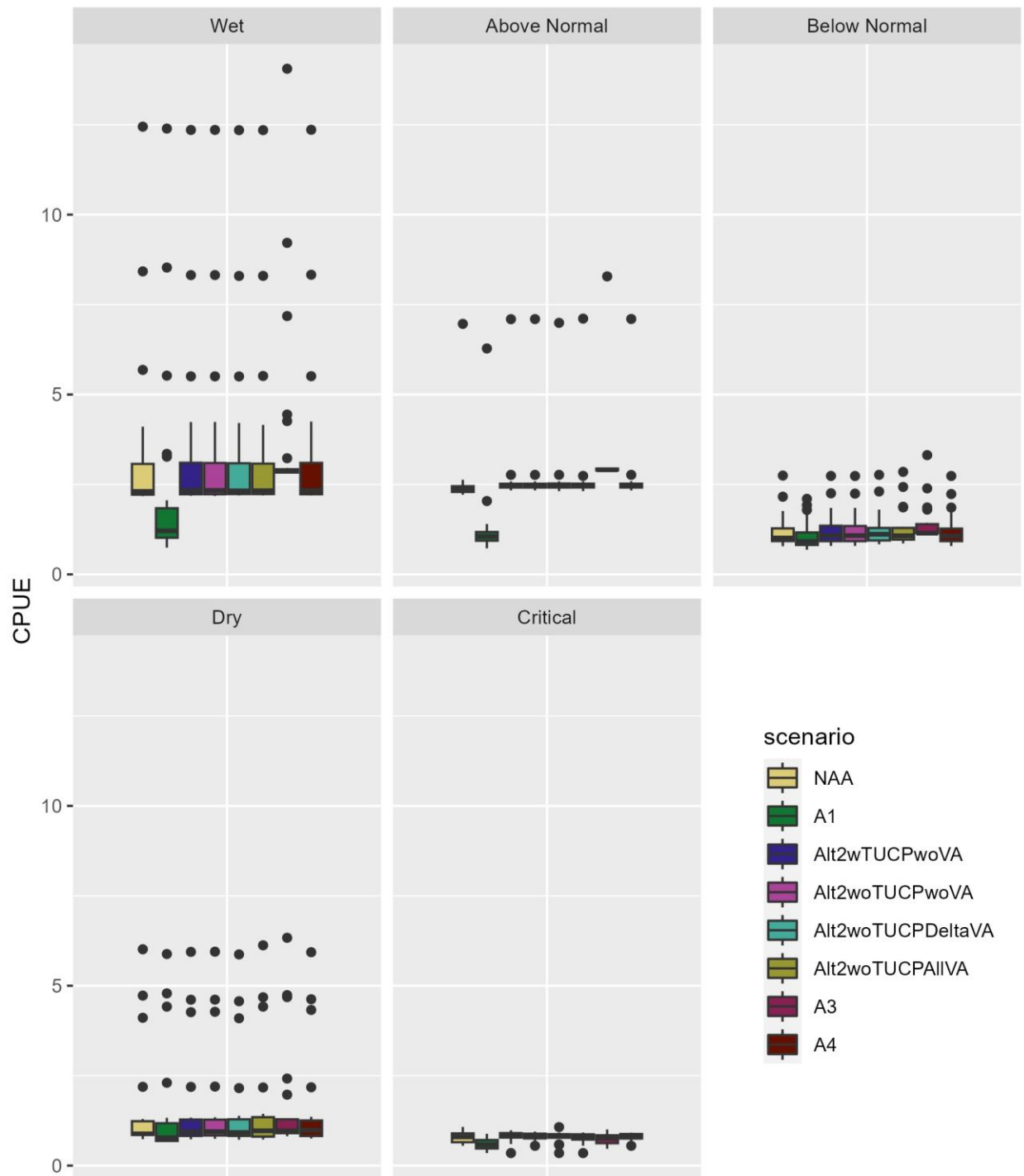


Figure J.3-16. Box Plots of CPUE of *E. affinis* adults by scenario across different water year types for fall.

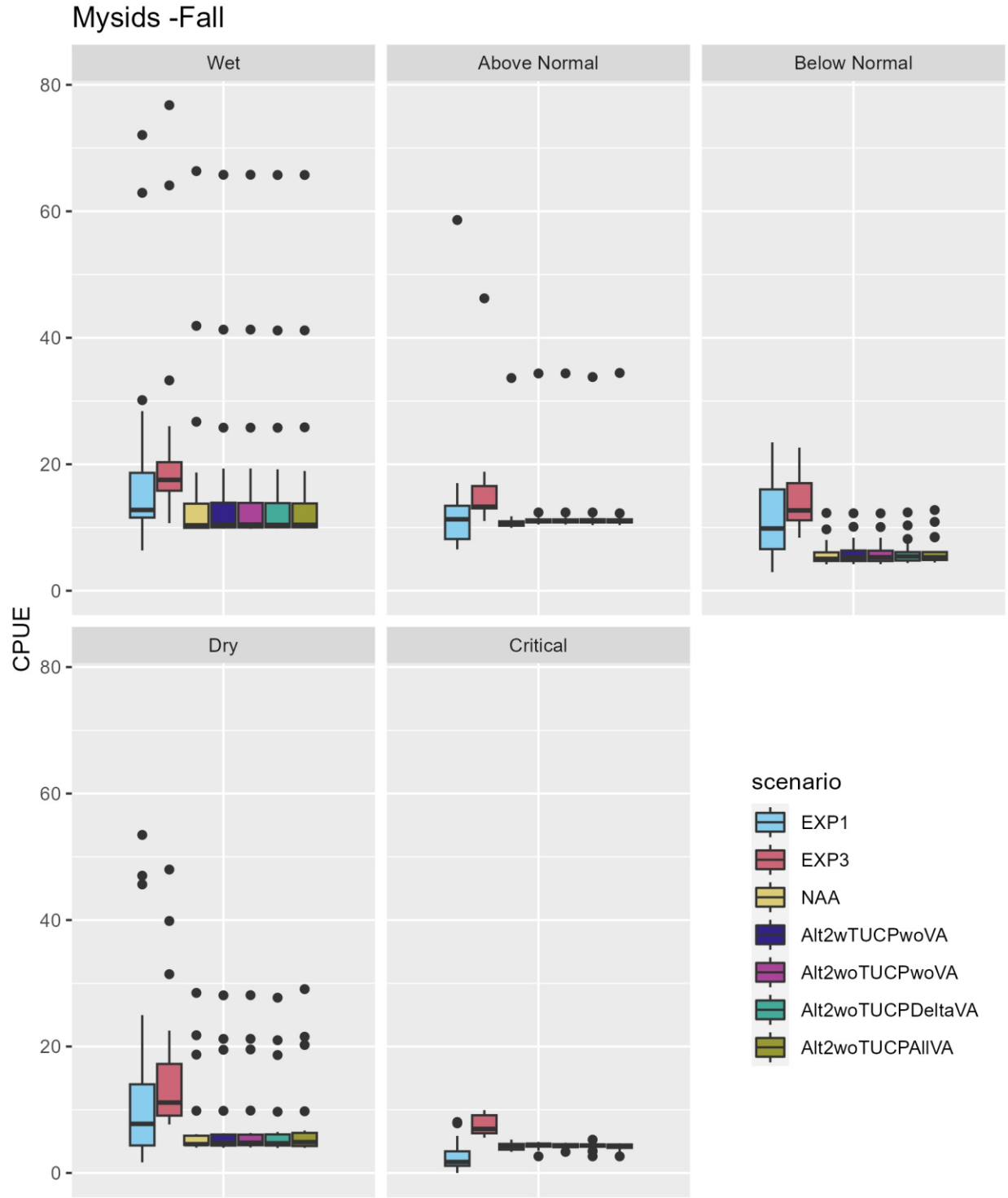


Figure J.3-17. Box Plots of CPUE of mysids by scenario across different water year types for fall.

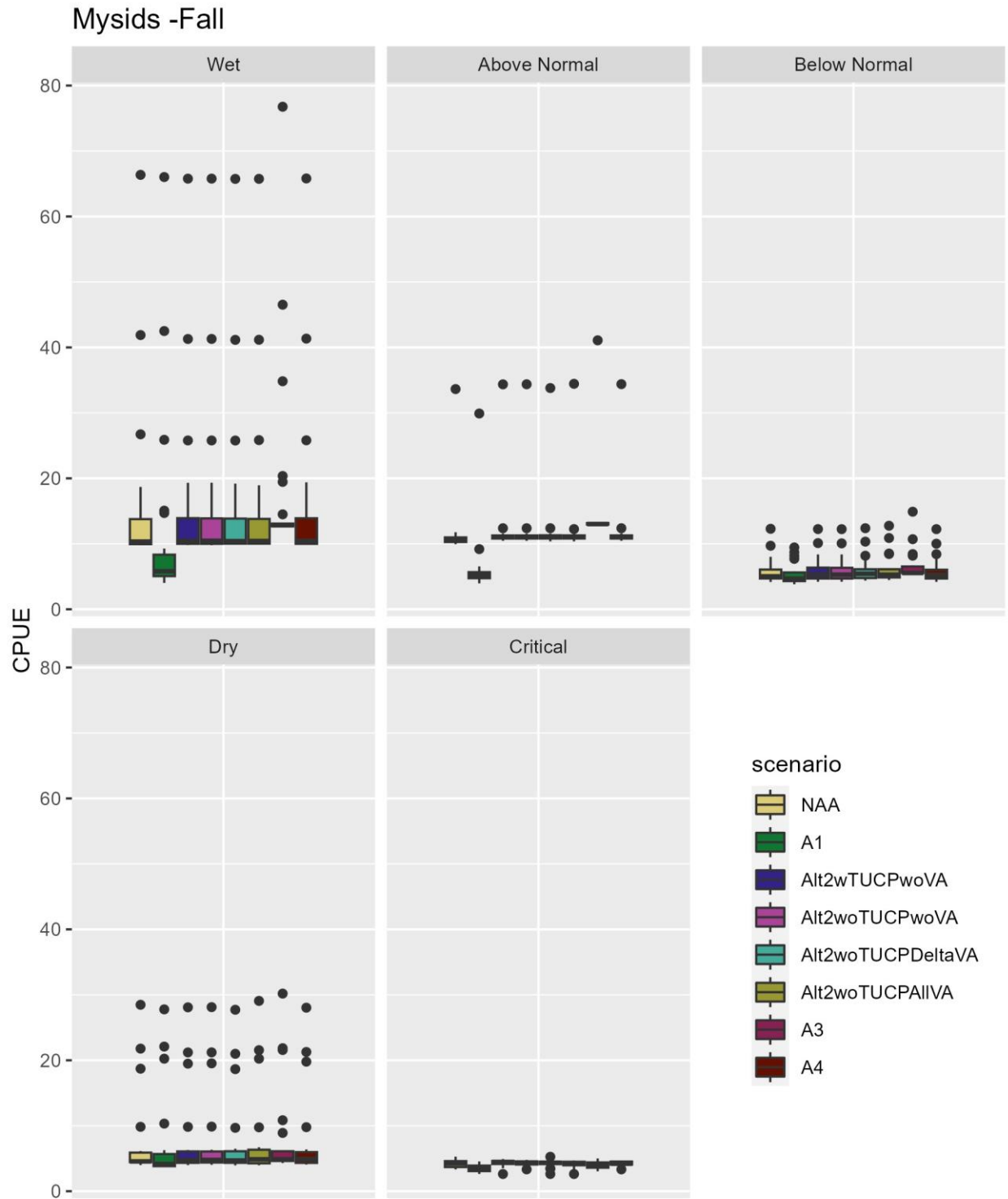


Figure J.3-18. Box Plots of CPUE of mysids by scenario across different water year types for fall.

J.3.4 References

Ambler et al. 1985

Barros, A., J. A. Hobbs, M. Willmes, C. M. Parker, M. Bisson, N. A. Fangué, A. L. Rypel, and L. S. Lewis. 2022. Spatial Heterogeneity in Prey Availability, Feeding Success, and Dietary Selectivity for the Threatened Longfin Smelt. *Estuaries and Coasts* 45:1766–1779.

Bashevkin, S. M., R. Hartman, K. Alstad, and C. Pien. 2023a. zooper: an R package to download and integrate zooplankton datasets from the Upper San Francisco Estuary. *Zenodo*. doi:10.5281/zenodo.3776867

Bashevkin, S. M., R. Hartman, M. Thomas, A. Barros, C. Burdi, A. Hennessy, T. Tempel, K. Kayfetz, K. Alstad, and C. Pien. 2023b. Interagency Ecological Program: Zooplankton abundance in the Upper San Francisco Estuary from 1972-2021, an integration of 7 long-term monitoring programs. Version 4. Environmental Data Initiative. doi:10.6073/pasta/8b646dfbeb625e308212Alt39f1e46f69b

Bashevkin, S. M., R. Hartman, M. Thomas, A. Barros, C. E. Burdi, A. Hennessy, T. Tempel, and K. Kayfetz. 2022. Five decades (1972–2020) of zooplankton monitoring in the upper San Francisco Estuary. *PLoS ONE* 17(3):e0265402.

Bollens, S. M., J. K. Breckenridge, J. R. Cordell, C. A. Simenstad, and O. Kalata. 2014. Zooplankton of tidal marsh channels in relation to environmental variables in the upper San Francisco Estuary. *Aquatic Biology* 21(3):205–219.

Brandon, J., C. Lee, A. Smith, S. Acuña, and A. Schultz. 2022. Detecting responses of Delta Smelt prey biomass to freshwater outflow management actions in a highly altered estuarine system: using power analysis to evaluate environmental monitoring sampling. Pages 239-266 in Bertrand, N. G., K. K. Arend, and B. Mahardja., editors. *Directed Outflow Project: Technical Report 3*. U.S. Bureau of Reclamation, Bay-Delta Office, California-Great Basin Region, Sacramento, CA. June 10, 2022.

Burris et al. 2022

California Department of Water Resources and U.S. Bureau of Reclamation. 2023. *Temporary Urgency Change Petition for February and March 2023*. Petition to State Water Resources Control Board. February 13.

Feyrer, F., B. Herbold, S. A. Matern, and P. B. Moyle. 2003. Dietary shifts in a stressed fish assemblage: consequences of a bivalve invasion in the San Francisco Estuary. *Environmental Biology of Fishes* 67:277–288.

Glibert, P. M., D. Fullerton, J. M. Burkholder, J. C. Cornwell, and T. M. Kana. 2011. Ecological stoichiometry, biogeochemical cycling, invasive species, and aquatic food webs: San Francisco Estuary and comparative systems. *Reviews in Fisheries Science* 19(4):358–417.

- Greenwood, M. 2018. *Potential Effects on Zooplankton from California WaterFix Operations*. Technical Memorandum to California Department of Water Resources. July 2. Available: https://www.waterboards.ca.gov/waterrights/water_issues/programs/bay_delta/california_waterfix/exhibits/docs/petitioners_exhibit/dwr/part2_rebuttal/dwr_1349.pdf. Accessed: February 15, 2022.
- Hassrick, J. L., J. Korman, W. J. Kimmerer, E. S. Gross, L. F. Grimaldo, C. Lee, and A. A. Schultz. 2023. Freshwater flow affects subsidies of a copepod (*Pseudodiaptomus forbesi*) to low-salinity food webs in the upper San Francisco estuary. *Estuaries and Coasts* 46(2):450–462.
- Hennessy, A., and Z. Burris. 2017. *Preliminary analysis of current relationships between zooplankton abundance and freshwater outflow in the upper San Francisco Estuary*. Memorandum to S. Louie, Senior Environmental Scientist, California Department of Fish and Wildlife, Water Branch. February 21. Stockton, CA: California Department of Fish and Wildlife, Bay-Delta Region.
- Kimmerer 2009
- Kimmerer, W. 2004. Open water processes of the San Francisco Estuary: from physical forcing to biological responses. *San Francisco Estuary and Watershed Science* 2(1).
- Kimmerer, W. J. 2002. Effects of Freshwater Flow on Abundance of Estuarine Organisms: Physical Effects or Trophic Linkages? *Marine Ecology Progress Series* 243:39–55.
- Kimmerer, W. J., E. S. Gross, A. M. Slaughter, and J. R. Durand. 2019. Spatial subsidies and mortality of an estuarine copepod revealed using a box model. *Estuaries and Coasts* 42:218–236.
- Kimmerer, W. J., J. R. Burau, and W. A. Bennett. 1998. Tidally oriented vertical migration and position maintenance of zooplankton in a temperate estuary. *Limnology and Oceanography* 43(7):1697–1709.
- Lee, C. Y., A. G. Smith, J. L. Hassrick, A. J. Kalmbach, M. C. Sabal, D. M. Cox, L. F. Grimaldo, and A. Schultz. 2023. Flow Augmentations Modify an Estuarine Prey Field. *San Francisco Estuary and Watershed Science* 21(2).
- Merz, J. E., P. S. Bergman, J. L. Simonis, D. Delaney, J. Pierson, and P. Anders. 2016. Long-term seasonal trends in the prey community of Delta Smelt (*Hypomesus transpacificus*) within the Sacramento-San Joaquin Delta, California. *Estuaries and Coasts* 39:1526–1536.
- Nobriga 2002
- R Core Team. 2023. *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing, Vienna, Austria. <https://www.R-project.org/>.
- Slater et al. 2019

- Slater, S. B., and R. D. Baxter. 2014. Diet, prey selection, and body condition of age-0 delta smelt, *Hypomesus transpacificus*, in the Upper San Francisco Estuary. *San Francisco Estuary and Watershed Science* 12(3).
- Smith, W. 2021. *A Delta Smelt Individual-Based Life Cycle Model in the R Statistical Environment*. 16 August.
- Smith, W. E., and M. L. Nobriga. 2023. A bioenergetics-based index of habitat suitability: spatial dynamics of foraging constraints and food limitation for a rare estuarine fish. *Transactions of the American Fisheries Society*. DOI: <http://dx.doi.org/10.1002/tafs.10427>.
- Thomson, J. R., W. J. Kimmerer, L. R. Brown, K. B. Newman, R. Mac Nally, W. A. Bennett, F. Feyrer, and E. Fleishman. 2010. Bayesian Change Point Analysis of Abundance Trends for Pelagic Fishes in the Upper San Francisco Estuary. *Ecological Applications* 20(5):1431–1448.
- Winder, M., and A. D. Jassby. 2011. Shifts in zooplankton community structure: implications for food web processes in the upper San Francisco Estuary. *Estuaries and Coasts* 34:675–690.